CASRN: 35822-46-9

Chemical Category: POLYCHLORINATED DIBENZO-p-DIOXINS

Chemical Name (Common Synonyms):

1,2,3,4,6,7,8-HEPTACHLORO DIBENZO-p-DIOXIN

Chemical Characteristics

Solubility in Water: No data [1], 2.4 mg/L [2] Half-Life: No data [2,3]

Log K_{ow}: No data [4], 8.00 [2] **Log K**_{oc}: 7.86 L/kg organic carbon

Human Health

Oral RfD: No data [5] Confidence: —

Critical Effect: —

Oral Slope Factor: No data [5] Carcinogenic Classification: —

Wildlife

Partitioning Factors: Partitioning factors for 1,2,3,4,6,7,8-heptaCDD in wildlife were not found in the studies reviewed.

Food Chain Multipliers: Limited information was found on specific biomagnification factors of PCDDs and PCDFs through terrestrial wildlife. Due to the toxicity, high K_{ow} values, and highly persistent nature of the PCDDs and PCDFs, they possess a high potential to bioaccumulate and biomagnify through the food web. PCDDs and PCDFs have been identified in fish and wildlife throughout the global aquatic and marine environments [6]. Studies conducted in Lake Ontario indicated that biomagnification of 2,3,7,8-tetrachlorodibenzo-*p*-dioxin (2,3,7,8-TCDD) appears to be significant between fish and fisheating birds but not between fish and their food. When calculated for older predaceous fish such as lake-trout-eating young smelt, the biomagnification factor (BMF) can equal 3. The log BMF from alewife to herring gulls in Lake Ontario was 1.51 for 2,3,7,8-TCDD [7]. Log BMFs of 1.18 to 1.70 were determined for mink fed 1,2,3,4,6,7,8-Hepta CDD in the diet [18].

EPA has developed risk-based concentrations of 2,3,7,8-TCDD in different media that present low and high risk to fish, mammalian, and avian wildlife. These concentrations were developed based on toxic effects of 2,3,7,8-TCDD and its propensity to bioaccumulate in fish, mammals, and birds.

Environmental Concentrations Associated With 2,3,7,8-TCDD Risk to Aquatic Life and Associated Wildlife [8]

	Fish Concentration	Sediment Concentration	Water Concentration (pg/L)						
Organism	(pg/g)	(pg/g dry wt.)	POC=0.2	POC=1.0					
Low Risk									
Fish	50	60	0.6	3.1					
Mammalian Wildlife	0.7	2.5	0.008	0.04					
Avian Wildlife	6	21	0.07	0.35					
	High R	isk to Sensitive Species							
Fish	80	100	1.0	5					
Mammalian Wildlife	7	25	0.08	0.4					
Avian Wildlife	60	210	0.7	3.5					

Note: POC - Particulate organic carbon

Fish lipid of 8% and sediment organic carbon of 3% assumed where needed.

For risk to fish, BSAF of 0.3 used; for risk to wildlife, BSAF of 0.1 used.

Low risk concentrations are derived from no-effects thresholds for reproductive effects (mortality in embryos and young) in sensitive species.

High risk concentrations are derived from TCDD doses expected to cause 50 to 100% mortality in embryos and young of sensitive species.

Aquatic Organisms

Partitioning Factors: In one study, the BSAF for carp collected from a reservoir in central Wisconsin was 0.0048. In a laboratory study, log BCFs for fathead minnow, rainbow trout, and goldfish exposed to 1,2,3,4,6,7,8-HeptaCDD were 2.71, 3.15, and 4.28, respectively.

Food Chain Multipliers: No specific food chain multipliers were identified for 1,2,3,4,6,7,8-heptaCDD. Food chain multiplier information was only available for 2,3,7,8-TCDD. Biomagnification of 2,3,7,8-TCDD does not appear to be significant between fish and their prey. Limited data for the base of the Lake Ontario lake trout food chain indicated little or no biomagnification between zooplankton and forage fish. BMFs based on fish consuming invertebrate species are probably close to 1.0 because of the 2,3,7,8-TCDD biotransformation by forage fish. BMFs greater than 1.0 may exist between some zooplankton species and their prey due to the lack of 2,3,7,8-TCDD biotransformation in invertebrates [8].

Toxicity/Bioaccumulation Assessment Profile

The polychlorinated dibenzo-*p*-dioxins (PCDDs) and polychlorinated dibenzofurans (PCDFs) each consist of 75 isomers that differ in the number and position of attached chlorine atoms. The PCDDs and PCDFs are polyhalogenated aromatic compounds and exhibit several properties common to this group of compounds. These compounds tend to be highly lipophilic and the degree of lipophilicity is increased with increasing ring chlorination [6]. In general, the PCDDs and PCDFs exhibit relative inertness to acids, bases, oxidation, reduction, and heat, increasing in environmental persistence and chemical

stability with increasing chlorination [9,6]. Because of their lipophilic nature, the PCDDs and PCDFs have been detected in fish, wildlife, and human adipose tissue, milk, and serum [6].

Each isomer has its own unique chemical and toxicological properties. The most toxic of the PCDD and PCDF isomers is 2,3,7,8-TCDD, one of the 22 possible congeners of tetrachlorodibenzo-*p*-dioxin [9]. Toxicity equivalency factors (TEFs) have been developed by EPA relating the toxicities of other PCDD and PCDF isomers to that of 2,3,7,8-TCDD [10]. The biochemical mechanisms leading to the toxic response resulting from exposure to PCDDs and PCDFs are not known in detail, but experimental data suggest that an important role in the development of systemic toxicity resulting from exposure to these chemicals is played by an intracellular protein, the Ah receptor. This receptor binds halogenated polycyclic aromatic molecules, including PCDDs and PCDFs. In several mouse strains, the expression of toxicity of 2,3,7,8-TCDD-related compounds, including cleft palate formation, liver damage, effects on body weight gain, thymic involution, and chloracnegenic response, has been correlated with their binding affinity for the Ah receptor, and with their ability to induce several enzyme systems [10].

Toxicity Equivalency Factors (TEF) for PCDD and PCDF Isomers [10]

Isomer	TEF
Total TetraCDD	1
2,3,7,8-TCDD	1
Other TCDDs	0.01
Total PentaCDDs	0.5
2,3,7,8-PentaCDDs	0.5
Other PentaCDDs	0.005
Total HexaCDDs	0.04
2,3,7,8-HexaCDDs	0.04
Other HexaCDDs	0.0004
Total HeptaCDDs	0.001
2,3,7,8-HeptaCDDs	0.001
Other HeptaCDDs	0.00001
Total TetraCDFs	0.1
2,3,7,8-TetraCDF	0.1
Other TetraCDFs	0.001
Total PentaCDFs	0.1
2,3,7,8-PentaCDFs	0.1
Other PentaCDFs	0.001
Total HexaCDFs	0.01
2,3,7,8-HexaCDFs	0.01
Other HexaCDFs	0.0001
Total HeptaCDFs	0.001
2,3,7,8-HeptaCDFs	0.001
Other HeptaCDFs	0.00001

In natural systems, PCDDs and PCDFs are typically associated with sediments, biota, and the organic carbon fraction of ambient waters [8]. Congener-specific analyses have shown that the 2,3,7,8-substituted PCDDs and PCDFs were the major compounds present in most sample extracts [6]. Results from limited epidemiology studies are consistent with laboratory-derived threshold levels to 2,3,7,8-TCDD impairment of reproduction in avian wildlife. Population declines in herring gulls (*Larus argentatus*) on Lake Ontario during the early 1970s coincided with egg concentrations of 2,3,7,8-TCDD and related chemicals expected to cause reproductive failure based on laboratory experiments (2,3,7,8-TCDD concentrations in excess of 1,000 pg/g). Improvements in herring gull reproduction through the mid-1980s were correlated with declining 2,3,7,8-TCDD concentrations in eggs and lake sediments [8]. Based on limited information on isomer-specific analysis from animals at different trophic levels, it appears that at higher trophic levels, i.e., fish-eating birds and fish, there is a selection of the planar congeners with the 2,3,7,8-substituted positions [11].

PCDDs and PCDFs are accumulated by aquatic organisms through exposure routes that are determined by the habitat and physiology of each species. With log K_{ow}>5, exposure through ingestion of contaminated food becomes an important route for uptake in comparison to respiration of water [8]. The relative contributions of water, sediment, and food to uptake of 2,3,7,8-TCDD by lake trout in Lake Ontario were examined by exposing yearling lake trout to Lake Ontario smelt and sediment from Lake Ontario along with water at a 2,3,7,8-TCDD concentration simulated to be at equilibrium with the sediments. Food ingestion was found to contribute approximately 75 percent of total 2,3,7,8-TCDD [8]. There have been a number of bioconcentration studies of 2,3,7,8-TCDD using model ecosystem and single species exposure. Although there is variation in the actual BCF values, in general, the algae and plants have the lowest BCF values, on the order of a few thousand. A log BCF value of 4.38 has been reported for the snail *Physa* sp. Crustacea and insect larvae appear to have the next highest BCF values, followed by several species of fish, with the highest log BCF value of 4.79 [11].

Exposure of juvenile rainbow trout to 2,3,7,8-TCDD and -TCDF in water for 28 days resulted in adverse effects on survival, growth, and behavior at extremely low concentrations. A no-observed-effects concentration (NOEC) for 2,3,7,8-TCDD could not be determined because the exposure to the lowest dose of 0.038 ng/l resulted in significant mortality [12]. A number of biological effects have been reported in fish following exposure to 2,3,7,8-TCDD including enzyme induction, immunological effects, wasting syndrome, dermatological effects, hepatic effects, hematological effects, developmental effects, and cardiovascular effects [11].

Species:	Concentra	tion, Units in ¹	:	Toxicity:	Ability t	o Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Fishes									
Salmonids							0.0031	[20]	F
Oncorhynchus mukiss (Salmo gairderni), Rainbow trout		Exposure water 11-55 ng/L			3.15 ± 2.35			[16]	L
Oncorhynchus mukiss (Salmo gairderni), Rainbow trout			0.000035 mg/kg (liver) ⁴	Biochemical, LOED				[19]	L; significant increase in liver ethoxyresorufin O-deethylase (EROD)
Cyprinus carpio, Carp	2,190 pg/g ⁵		27 pg/g ⁵				0.0048	[13]	F; Petenwell Reservoir, central Wisconsin; BSAF based on 8% tissue lipid content and 3.1% sediment organic carbon

Species:	Concentration, Units in ¹ :		:	Toxicity:	Ability to Accumulate ² :			Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Carassius auratus, Goldfish			1.91/2.2 ng/g ⁵ (whole body)		4.28			[15]	L; fish were exposed for 120 hr; exposure water contained fly ash extract; concentrations were measured in water, but data were not presented
Pimephales promelas, Fathead minnow		Exposure water 8-39 ng/L			2.71 ± 2.03			[16]	L
Platycephalus caerulopunctatus and Platycephalus bassensis, Flathead	0.356 pg/g, dw		558 pg/kg					[14]	F; unimpacted coastal site; surface sediment composite; most other dioxin congeners were below
Sillago bassensis, School whiting	0.356 pg/g, dw		375 pg/kg						detection.
Wildlife									
Falco peregrinus, Peregrine falcon			0.7 ng/g (eggs) $(n = 6)$	11.4% eggshell thinning				[17]	F; Kola Peninsula, Russia

Species:	Concentra	tion, Units in	:	Toxicity:	Ability to Accumulate ² :			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Mustela vison, Mink	Diet: 5 pg/g ⁵		115 pg/g ⁵ (liver)	NOAEL		log BMF = 1.18		[18]	L; BMF = lipid-normalized concentration in the liver	
	7 pg/g ⁵		330 pg/g ⁵ (liver)	LOAEL; reduced kit body weights followed by reduced survival		log BMF = 1.70			divided by the lipid-normalized dietary concentration	
	6 pg/g ⁵		290 pg/g ⁵ (liver)	Reduced kit body weights followed by reduced survival		log BMF = 1.69				
	13 pg/g ⁵		380 pg/g ⁵ (liver)	Significant decrease in number of live kits whelped per female		log BMF = 1.66				

¹ Concentration units based on wet weight unless otherwise noted.

² BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ This entry was excerpted directly from the Environmental Residue-Effects Database (ERED, www.wes.army.mil/el/ered, U.S. Army Corps of Engineers and U.S. Environmental Protection Agency). The original publication was not reviewed, and the reader is strongly urged to consult the publication to confirm the information presented here.

⁵ Not clear from reference if concentration is based on wet or dry weight.

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Chemical Category: POLYCHLORINATED DIBENZO-p-DIOXINS

Chemical Name (Common Synonyms): CASRN: 39227-28-6

1,2,3,4,7,8-HEXACHLORODIBENZO-p-DIOXIN

Chemical Characteristics

Solubility in Water: No data [1], **Half-Life:** No data [2,3]

 $8.25 \times 10^{-6} \text{ mg/L } [1,2]$

Log K_{ow}: No data [4], 7.70 [2] **Log K_{oc}:** 7.57 L/kg organic carbon

Human Health

Oral RfD: No data [5] Confidence: —

Critical Effect: —

Oral Slope Factor (Reference): No data [5] Carcinogenic Classification: —

Wildlife

Partitioning Factors: Partitioning factors for 1,2,3,4,7,8-hexaCDD in wildlife were not found in the studies reviewed.

Food Chain Multipliers: Limited information was found reporting on specific biomagnification factors of PCDDs and PCDFs through terrestrial wildlife. Due to the toxicity, high K_{ow} values, and highly persistent nature of the PCDDs and PCDFs, they possess a high potential to bioaccumulate and biomagnify through the food web. PCDDs and PCDFs have been identified in fish and wildlife throughout the global aquatic and marine environments [6]. Studies conducted in Lake Ontario indicated that biomagnification of 2,3,7,8-tetrachlorodibenzo-p-dioxin (2,3,7,8-TCDD) appears to be significant between fish and fish-eating birds but not between fish and their food. When calculated for older predaceous fish such as lake-trout-eating young smelt, the biomagnification factor (BMF) can equal 3. The BMF from alewife to herring gulls in Lake Ontario was 32 for 2,3,7,8-TCDD [7]. A log BMF of 0.97 was reported for mink exposed to 1,2,3,4,7,8-hexaCDD in the diet. [14].

EPA has developed risk-based concentrations of 2,3,7,8-TCDD in different media that present low and high risk to fish, mammalian, and avian wildlife. These concentrations were developed based on toxic effects of 2,3,7,8-TCDD and its propensity to bioaccumulate in fish, mammals, and birds.

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Note: POC - Particulate organic carbon

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For risk to fish, BSAF of 0.3 used; for risk to wildlife, BSAF of 0.1 used.

Low risk concentrations are derived from no-effects thresholds for reproductive effects (mortality in embryos and young) in sensitive species.

High risk concentrations are derived from TCDD doses expected to cause 50 to 100% mortality in embryos and young of sensitive species.

Aquatic Organisms

Partitioning Factors: In a laboratory study, log BCFs for rainbow trout and fathead minnow exposed to 1,2,3,4,7,8-HexaCDD were 3.73 and 4.00, respectively.

Food Chain Multipliers: No specific food chain multipliers were identified for 1,2,3,4,7,8-hexaCDD. Food chain multiplier information was only available for 2,3,7,8-TCDD. Biomagnification of 2,3,7,8-TCDD does not appear to be significant between fish and their prey. Limited data for the base of the Lake Ontario lake trout food chain indicated little or no biomagnification between zooplankton and forage fish. BMFs based on fish consuming invertebrate species are probably close to 1.0 because of the 2,3,7,8-TCDD biotransformation by forage fish. BMFs greater than 1.0 may exist between some zooplankton species and their prey due to the lack of 2,3,7,8-TCDD biotransformation in invertebrates [8].

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Other TCDDs	0.01
Total PentaCDDs	0.5
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Other PentaCDDs	0.005
Total HexaCDDs	0.04
2,3,7,8-HexaCDDs	0.04
Other HexaCDDs	0.0004
Total HeptaCDDs	0.001
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Other HeptaCDDs	0.00001
Total TetraCDFs	0.1
2,3,7,8-TetraCDF	0.1
Other TetraCDFs	0.001
Total PentaCDFs	0.1
2,3,7,8-PentaCDFs	0.1
Other PentaCDFs	0.001
Total HexaCDFs	0.01
2,3,7,8-HexaCDFs	0.01
Other HexaCDFs	0.0001
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Species:	Species: Concentration, Units in ¹ :			Toxicity:	Ability	to Accum	ulate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Fishes									
Oncorhynchus mukiss (Salmo gairderni), Rainbow trout		Exposure water 10-47 ng/L			3.73			[11]	L
Oncorhynchus mukiss, Rainbow trout			0.0000395 mg/kg (liver) ⁴	Biochemical, LOED				[15]	L; significant increase in liver ethoxyresorufin O-deethylase (EROD)
Pimephales promelas, Fathead minnow		Exposure water 10-47 ng/L			4.00			[11]	L
Wildlife									
Falco peregrinus, Peregrine falcon			3.3 ng/g (eggs) (n = 6)	11.4% eggshell thinning				[13]	F; Kola Peninsula, Russia

Species:	Concentration, Units in ¹ :		Toxicity:	Ability to Accumulate ² :			Source:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Mustela vison, Mink	Diet: 2 pg/g ⁵		6 pg/g ⁵ (liver)	LOAEL; reduced kit body weights followed by reduced survival		No BMF reported		[14]	L; BMF = biomagnification factor = v_1/v_d v_1 = lipid-normalized tissue concentration,
	1 pg/g ⁵		77 pg/g ⁵ (liver)	Reduced kit body weights followed by reduced survival		No BMF reported			v _d = lipid- normalized dietary concentration.
	3 pg/g ⁵		15 pg/g ⁵ (liver)	Significant decrease in number of live kits whelped per female		Log BMF = 0.97			

¹ Concentration units based on wet weight unless otherwise noted.

BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ This entry was excerpted directly from the Environmental Residue-Effects Database (ERED, www.wes.army.mil/el/ered, U.S. Army Corps of Engineers and U.S. Environmental Protection Agency). The original publication was not reviewed, and the reader is strongly urged to consult the publication to confirm the information presented here.

Not clear from reference if concentration is based on wet or dry weight.

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CASRN: 57653-85-7

Chemical Category: POLYCHLORINATED DIBENZO-p-DIOXINS

Chemical Name (Common Synonyms):

1,2,3,6,7,8-HEXACHLORODIBENZO-p-DIOXIN

Chemical Characteristics

Solubility in Water: No data [1,2] **Half-Life:** No data [2,3]

Log K_{ow}: No data [2,4] **Log K**_{oc}: —

Human Health

Oral RfD: No data [5] Confidence: —

Critical Effect: Hepatic tumors in mice and rats

Oral Slope Factor: 6.2 x 10⁺³ per (mg/kg)/day [5] Carcinogenic Classification: B2 [5]

Wildlife

Partitioning Factors: Partitioning factors for 1,2,3,6,7,8-hexaCDD in wildlife were not found in the studies reviewed.

Food Chain Multipliers: Limited information was found reporting on specific biomagnification factors of PCDDs and PCDFs through terrestrial wildlife. Due to the toxicity, high K_{ow} values, and highly persistent nature of the PCDDs and PCDFs, they possess a high potential to bioaccumulate and biomagnify through the food web. PCDDs and PCDFs have been identified in fish and wildlife throughout the global aquatic and marine environments [6]. Studies conducted in Lake Ontario indicated that biomagnification of 2,3,7,8-tetrachlorodibenzo-p-dioxin (2,3,7,8-TCDD) appears to be significant between fish and fish-eating birds but not between fish and their food. When calculated for older predaceous fish such as lake-trout-eating young smelt, the biomagnification factor (BMF) can equal 3. The BMF from alewife to herring gulls in Lake Ontario was 32 for 2,3,7,8-TCDD [7]. Log BMFs of 1.42 and 1.43 were reported for mink exposed to 1,2,3,6,7,8-hexaCDD in the diet [18].

EPA has developed risk-based concentrations of 2,3,7,8-TCDD in different media that present low and high risk to fish, mammalian, and avian wildlife. These concentrations were developed based on toxic effects of 2,3,7,8-TCDD and its propensity to bioaccumulate in fish, mammals, and birds.

Environmental Concentrations Associated With 2,3,7,8-TCDD Risk to Aquatic Life and Associated Wildlife [8]

	Eigh Component of in	Sediment	Water Concentration (pg/L)						
Organism	Fish Concentration (pg/g)	Concentration (pg/g dry wt.)	POC=0.2	POC=1.0					
Low Risk									
Fish	50	60	0.6	3.1					
Mammalian Wildlife	0.7	2.5	0.008	0.04					
Avian Wildlife	6	21	0.07	0.35					
	High R	isk to Sensitive Species							
Fish	80	100	1.0	5					
Mammalian Wildlife	7	25	0.08	0.4					
Avian Wildlife	60	210	0.7	3.5					

Note: POC - Particulate organic carbon

Fish lipid of 8% and sediment organic carbon of 3% assumed where needed.

For risk to fish, BSAF of 0.3 used; for risk to wildlife, BSAF of 0.1 used.

Low risk concentrations are derived from no-effects thresholds for reproductive effects (mortality in embryos and young) in sensitive species.

High risk concentrations are derived from TCDD doses expected to cause 50 to 100% mortality in embryos and young of sensitive species.

Aquatic Organisms

Partitioning Factors: In one study, the BSAF for carp collected from a reservoir in central Wisconsin was 0.035. The log BCF for goldfish during a laboratory exposure for 120 hours was 4.61.

Food Chain Multipliers: No specific food chain multipliers were identified for 1,2,3,6,7,8-hexaCDD. Food chain multiplier information was only available for 2,3,7,8-TCDD. Biomagnification of 2,3,7,8-TCDD does not appear to be significant between fish and their prey. Limited data for the base of the Lake Ontario lake trout food chain indicated little or no biomagnification between zooplankton and forage fish. BMFs based on fish consuming invertebrate species are probably close to 1.0 because of the 2,3,7,8-TCDD biotansformation by forage fish. BMFs greater than 1.0 may exist between some zooplankton species and their prey due to the lack of 2,3,7,8-TCDD biotransformation in invertebrates[8].

Toxicity/Bioaccumulation Assessment Profile

The polychlorinated dibenzo-*p*-dioxins (PCDDs) and polychlorinated dibenzofurans (PCDFs) each consist of 75 isomers that differ in the number and position of attached chlorine atoms. The PCDDs and PCDFs are polyhalogenated aromatic compounds and exhibit several properties common to this group of compounds. These compounds tend to be highly lipophilic and the degree of lipophilicity is increased with increasing ring chlorination [6]. In general, the PCDDs and PCDFs exhibit relative inertness to acids, bases, oxidation, reduction, and heat, increasing in environmental persistence and chemical

stability with increasing chlorination [9,6]. Because of their lipophilic nature, the PCDDs and PCDFs have been detected in fish, wildlife, and human adipose tissue, milk, and serum [6].

Each isomer has its own unique chemical and toxicological properties. The most toxic of the PCDD and PCDF isomers is 2,3,7,8 TCDD, one of the 22 possible congeners of tetrachlorodibenzo-*p*-dioxin [9]. Toxicity equivalency factors (TEFs) have been developed by EPA relating the toxicities of other PCDD and PCDF isomers to that of 2,3,7,8-TCDD [10]. The biochemical mechanisms leading to the toxic response resulting from exposure to PCDDs and PCDFs are not known in detail, but experimental data suggest that an important role in the development of systemic toxicity resulting from exposure to these chemicals is played by an intracellular protein, the Ah receptor. This receptor binds halogenated polycyclic aromatic molecules, including PCDDs and PCDFs. In several mouse strains, the expression of toxicity of 2,3,7,8-TCDD-related compounds, including cleft palate formation, liver damage, effects on body weight gain, thymic involution, and chloracnegenic response, has been correlated with their binding affinity for the Ah receptor, and with their ability to induce several enzyme systems [10].

Toxicity Equivalency Factors (TEF) for PCDD and PCDF Isomers [10]

Isomer	TEF
Total TetraCDD	1
2,3,7,8-TCDD	1
Other TCDDs	0.01
Total PentaCDDs	0.5
2,3,7,8-PentaCDDs	0.5
Other PentaCDDs	0.005
Total HexaCDDs	0.04
2,3,7,8-HexaCDDs	0.04
Other HexaCDDs	0.0004
Total HeptaCDDs	0.001
2,3,7,8-HeptaCDDs	0.001
Other HeptaCDDs	0.00001
Total TetraCDFs	0.1
2,3,7,8-TetraCDF	0.1
Other TetraCDFs	0.001
Total PentaCDFs	0.1
2,3,7,8-PentaCDFs	0.1
Other PentaCDFs	0.001
Total HexaCDFs	0.01
2,3,7,8-HexaCDFs	0.01
Other HexaCDFs	0.0001
Total HeptaCDFs	0.001
2,3,7,8-HeptaCDFs	0.001
Other HeptaCDFs	0.00001

In natural systems, PCDDs and PCDFs are typically associated with sediments, biota, and the organic carbon fraction of ambient waters [7]. Congener-specific analyses have shown that the 2,3,7,8-substituted PCDDs and PCDFs were the major compounds present in most sample extracts [6]. Results from limited epidemiology studies are consistent with laboratory-derived threshold levels to 2,3,7,8-TCDD impairment of reproduction in avian wildlife. Population declines in herring gulls (*Larus argentatus*) on Lake Ontario during the early 1970s coincided with egg concentrations of 2,3,7,8-TCDD and related chemicals expected to cause reproductive failure based on laboratory experiments (2,3,7,8-TCDD concentrations in excess of 1,000 pg/g). Improvements in herring gull reproduction through the mid-1980s were correlated with declining 2,3,7,8-TCDD concentrations in eggs and lake sediments [8]. Based on limited information on isomer-specific analysis from animals at different trophic levels, it appears that at higher trophic levels, i.e., fish-eating birds and fish, there is a selection of the planar congeners with the 2,3,7,8-substituted positions [11].

PCDDs and PCDFs are accumulated by aquatic organisms through exposure routes that are determined by the habitat and physiology of each species. With log K_{ow}>5, exposure through ingestion of contaminated food becomes an important route for uptake in comparison to respiration of water [8]. The relative contributions of water, sediment, and food to uptake of 2,3,7,8-TCDD by lake trout in Lake Ontario were examined by exposing yearling lake trout to Lake Ontario smelt and sediment from Lake Ontario along with water at a 2,3,7,8-TCDD concentration simulated to be at equilibrium with the sediments. Food ingestion was found to contribute approximately 75 percent of total 2,3,7,8-TCDD [8]. There have been a number of bioconcentration studies of 2,3,7,8-TCDD using model ecosystem and single species exposure. Although there is variation in the actual BCF values, in general, the algae and plants have the lowest BCF values, on the order of a few thousand. A log value of 4.38 has been reported for the snail *Physa* sp. Crustacea and insect larvae appear to have the next highest BCF values, followed by several species of fish, with the highest log BCF value of 4.79 [11].

Exposure of juvenile rainbow trout to 2,3,7,8-TCDD and -TCDF in water for 28 days resulted in adverse effects on survival, growth, and behavior at extremely low concentrations. A no-observed-effects concentration (NOEC) for 2,3,7,8-TCDD could not be determined because the exposure to the lowest dose of 0.038 ng/l resulted in significant mortality [12]. A number of biological effects have been reported in fish following exposure to 2,3,7,8-TCDD including enzyme induction, immunological effects, wasting syndrome, dermatological effects, hepatic effects, hematological effects, developmental effects, and cardiovascular effects [11].

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	e Comments ³
Fishes									
Carassius auratus, Goldfish			0.79 ng/g ⁴ (whole body)		4.61			[14]	L; fish were exposed for 120 hr; exposure water contained fly ash extract; concen- trations were measured in water, but data were not presented
Cyprinus carpio, Carp	180 pg/g ⁴		16 pg/g ⁴				0.035	[13]	F; Petenwell Reservoir, central Wisconsin; BSAF based on 8% tissue lipid content and 3.1% sediment organic carbon
Salmonids							0.0073	[19]	F

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability to	o Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	e Comments ³
Wildlife									
Falco peregrinus, Peregrine falcon			7.2 ng/g (eggs) (n = 6)	11.4% eggshell thinning				[17]	F; Kola Peninsula, Russia
Haliaeetus leucocephalus, Bald eagle chicks			Powell River site: ~9,000 ng/kg lipid weight basis (yolk sac) Reference site: ~500 ng/kg lipid weight basis (yolk sac)	A hepatic cytochrome P4501A cross-reactive protein (CYP1A) was induced nearly 6-fold in chicks from Powell River site compared to the reference (p < 0.05). No significant concentration-related effects were found for morphological, physiological, or histological parameters.				[15]	F; southern coast of British Columbia; eggs were collected from nests and hatched in the lab; ~ indicates value was taken from a figure

Species:	Concentrati	ion, Units in¹:		Toxicity:	Ability t	o Accumul	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	e Comments ³
Ardea herodias, Great blue heron chicks		Nicomekl site: 10±3.4 [16 ng/kg (egg) (n=11)	[16]	L; eggs were collected from three British Columbia					
			Vancouver site: 89±45.4 ng/kg (egg) (n=12)	Depression of growth compared to Nicomekl site. Presence of edema.				diffe cont incu	colonies with different levels of contamination and incubated in the laboratory
			Crofton site: 430±105.9 ng/kg (egg) (n=6)	Depression of growth compared to Nicomekl site. Presence of edema.					

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumula	te ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Referenc	e Comments ³
Mustela vison, Mink	Diet: 1 pg/g ⁴		54 pg/g ⁴ (liver)	LOAEL; reduced kit body weights followed by reduced survival		$\begin{array}{lll} \text{No BMF} & \text{biomag} \\ \text{reported} & \text{factor} = \\ v_1 = \text{lipi} \\ \text{normali} \\ \text{concent} \\ v_d = \text{lip} \end{array}$	L; BMF = biomagnification factor = v_l/v_d , v_l = lipid-normalized tissue concentration, v_d = lipid-normalized dietary		
	3 pg/g ⁴		77 pg/g ⁴ (liver)	Reduced kit body weights followed by reduced survival		log BMF = 1.42			concentration.
	6 pg/g ⁴		130 pg/g ⁴ (liver)	Significant decrease in number of live kits whelped per female		log BMF = 1.53			

¹ Concentration units based on wet weight unless otherwise noted.
² BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ Not clear from reference if concentration is based on wet or dry weight.

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CASRN: 40321-76-4

Chemical Category: POLYCHLORINATED DIBENZO-p-DIOXINS

Chemical Name (Common Synonyms):

1,2,3,7,8-PENTACHLORODIBENOZ-p-DIOXIN

Chemical Characteristics

Solubility in Water: No data [1,3] **Half-Life:** No data [2,3]

Log K_{ow}: No data [3,4] **Log K**_{oc}: —

Human Health

Oral RfD: No data [5] Confidence: —

Critical Effect: —

Oral Slope Factor: No data [5] Carcinogenic Classification: —

Wildlife

Partitioning Factors: Partitioning factors for 1,2,3,7,8-pentaCDD in wildlife were not found in the studies reviewed.

Food Chain Multipliers: Limited information was found reporting on specific biomagnification factors of PCDDs and PCDFs through terrestrial wildlife; no information was available for 1,2,3,7,8-pentaCDD, specifically. Due to the toxicity, high K_{ow} values, and highly persistent nature of the PCDDs and PCDFs, they possess a high potential to bioaccumulate and biomagnify through the food web. PCDDs and PCDFs have been identified in fish and wildlife throughout the global aquatic and marine environments [6]. Studies conducted in Lake Ontario indicated that biomagnification of 2,3,7,8-tetrachlorodibenzo-p-dioxin (2,3,7,8-TCDD) appears to be significant between fish and fish-eating birds but not between fish and their food. When calculated for older predaceous fish such as lake-trout-eating young smelt, the log biomagnification factor (BMF) can equal 0.48. The log BMF from alewife to herring gulls in Lake Ontario was 1.51 for 2,3,7,8-TCDD [7].

EPA has developed risk-based concentrations of 2,3,7,8-TCDD in different media that present low and high risk to fish, mammalian, and avian wildlife. These concentrations were developed based on toxic effects of 2,3,7,8-TCDD and its propensity to bioaccumulate in fish, mammals, and birds.

Environmental Concentrations Associated With 2,3,7,8-TCDD Risk to Aquatic Life and Associated Wildlife [8]

	Fish Concentration	Sediment Concentration	Water Concentration (pg/L)							
Organism	(pg/g)	(pg/g dry wt.)	POC=0.2	POC=1.0						
Low Risk										
Fish	50	60	0.6	3.1						
Mammalian Wildlife	0.7	2.5	0.008	0.04						
Avian Wildlife	6	21	0.07	0.35						
	High Risk to Sensitive Species									
Fish	80	100	1.0	5						
Mammalian Wildlife	7	25	0.08	0.4						
Avian Wildlife	60	210	0.7	3.5						

Note: POC - Particulate organic carbon

Fish lipid of 8% and sediment organic carbon of 3% assumed where needed.

For risk to fish, BSAF of 0.3 used; for risk to wildlife, BSAF of 0.1 used.

Low risk concentrations are derived from no-effects thresholds for reproductive effects (mortality in embryos and young) in sensitive species.

High risk concentrations are derived from TCDD doses expected to cause 50 to 100% mortality in embryos and young of sensitive species.

Aquatic Organisms

Partitioning Factors: Partitioning factors for 1,2,3,7,8-pentaCDF in aquatic organisms were not found in the studies reviewed.

Food Chain Multipliers: No specific food chain multipliers were identified for 1,2,3,7,8-pentaCDD. Food chain multiplier information was only available for 2,3,7,8-TCDD. Biomagnification of 2,3,7,8-TCDD does not appear to be significant between fish and their prey. Limited data for the base of the Lake Ontario lake trout food chain indicated little or no biomagnification between zooplankton and forage fish. BMFs based on fish consuming invertebrate species are probably close to 1.0 because of the 2,3,7,8-TCDD biotransformation by forage fish. BMFs greater than 1.0 might exist between some zooplankton species and their prey due to the lack of 2,3,7,8-TCDD biotransformation in invertebrates[8].

Toxicity/Bioaccumulation Assessment Profile

The polychlorinated dibenzo-*p*-dioxins (PCDDs) and polychlorinated dibenzofurans (PCDFs) each consist of 75 isomers that differ in the number and position of attached chlorine atoms. The PCDDs and PCDFs are polyhalogenated aromatic compounds and exhibit several properties common to this group of compounds. These compounds tend to be highly lipophilic and the degree of lipophilicity is increased with increasing ring chlorination [6]. In general, the PCDDs and PCDFs exhibit relative inertness to acids, bases, oxidation, reduction, and heat, increasing in environmental persistence and chemical stability with increasing chlorination [6,9]. Because of their lipophilic nature, the PCDDs and PCDFs have been detected in fish, wildlife, and human adipose tissue, milk, and serum [6].

Each isomer has its own unique chemical and toxicological properties. The most toxic of the PCDD and PCDF isomers is 2,3,7,8-TCDD, one of the 22 possible congeners of tetrachlorodibenzo-*p*-dioxin [9]. Toxicity equivalency factors (TEFs) have been developed by the U.S. EPA relating the toxicities of other PCDD and PCDF isomers to that of 2,3,7,8-TCDD [10]. The biochemical mechanisms leading to the toxic response resulting from exposure to PCDDs and PCDFs are not known in detail, but experimental data suggest that an important role in the development of systemic toxicity resulting from exposure to these chemicals is played by an intracellular protein, the Ah receptor. This receptor binds halogenated polycyclic aromatic molecules, including PCDDs and PCDFs. In several mouse strains, the expression of toxicity of 2,3,7,8-TCDD-related compounds, including cleft palate formation, liver damage, effects on body weight gain, thymic involution, and chloracnegenic response, has been correlated with their binding affinity for the Ah receptor, and with their ability to induce several enzyme systems [10].

Toxicity Equivalency Factors (TEF) for PCDD and PCDF Isomers [10]

Isomer	TEF
Total TetraCDD	1
2,3,7,8-TCDD	1
Other TCDDs	0.01
Total PentaCDDs	0.5
2,3,7,8-PentaCDDs	0.5
Other PentaCDDs	0.005
Total HexaCDDs	0.04
2,3,7,8-HexaCDDs	0.04
Other HexaCDDs	0.0004
Total HeptaCDDs	0.001
2,3,7,8-HeptaCDDs	0.001
Other HeptaCDDs	0.00001
Total TetraCDFs	0.1
2,3,7,8-TetraCDF	0.1
Other TetraCDFs	0.001
Total PentaCDFs	0.1
2,3,7,8-PentaCDFs	0.1
Other PentaCDFs	0.001
Total HexaCDFs	0.01
2,3,7,8-HexaCDFs	0.01
Other HexaCDFs	0.0001
Total HeptaCDFs	0.001
2,3,7,8-HeptaCDFs	0.001
Other HeptaCDFs	0.00001

In natural systems, PCDDs and PCDFs are typically associated with sediments, biota, and the organic carbon fraction of ambient waters [8]. Congener-specific analyses have shown that the 2,3,7,8-substituted PCDDs and PCDFs were the major compounds present in most sample extracts [6]. Results

from limited epidemiology studies are consistent with laboratory-derived threshold levels to 2,3,7,8-TCDD impairment of reproduction in avian wildlife. Population declines in herring gulls (*Larus argentatus*) on Lake Ontario during the early 1970s coincided with egg concentrations of 2,3,7,8-TCDD and related chemicals expected to cause reproductive failure based on laboratory experiments (2,3,7,8-TCDD concentrations in excess of 1,000 pg/g). Improvements in herring gull reproduction through the mid-1980s were correlated with declining 2,3,7,8-TCDD concentrations in eggs and lake sediments [8]. Based on limited information on isomer-specific analysis from animals at different trophic levels, it appears that at higher trophic levels, i.e., fish-eating birds and fish, there is a selection of the planar congeners with the 2,3,7,8-substituted positions [11].

PCDDs and PCDFs are accumulated by aquatic organisms through exposure routes that are determined by the habitat and physiology of each species. With log K_{ow}>5, exposure through ingestion of contaminated food becomes an important route for uptake in comparison to respiration of water [8]. The relative contributions of water, sediment, and food to uptake of 2,3,7,8-TCDD by lake trout in Lake Ontario were examined by exposing yearling lake trout to Lake Ontario smelt and sediment from Lake Ontario along with water at a 2,3,7,8-TCDD concentration simulated to be at equilibrium with the sediments. Food ingestion was found to contribute approximately 75 percent of total 2,3,7,8-TCDD [8]. There have been a number of bioconcentration studies of 2,3,7,8-TCDD using model ecosystem and single species exposure. Although there is variation in the actual log BCF values, in general, the algae and plants have the lowest BCF values, on the order of a few thousand. A log BCF value of 4.38 has been reported for the snail *Physa* sp. Crustacea and insect larvae appear to have the next highest BCF values, followed by several species of fish, with the highest log BCF value of 4.78 [11].

Exposure of juvenile rainbow trout to 2,3,7,8-TCDD and -TCDF in water for 28 days resulted in adverse effects on survival, growth, and behavior at extremely low concentrations. A no-observed-effects concentration (NOEC) for 2,3,7,8-TCDD could not be determined because the exposure to the lowest dose of 0.038 ng/l resulted in significant mortality [12]. A number of biological effects have been reported in fish following exposure to 2,3,7,8-TCDD including enzyme induction, immunological effects, wasting syndrome, dermatological effects, hepatic effects, hematological effects, developmental effects, and cardiovascular effects [11].

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability t	o Accumul	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Fishes									
Carassius auratus, Goldfish			1.59/2.61 ng/g ⁴ (whole body)		16,982			[14]	L; fish were exposed for 120 hr; exposure water contained fly ash extract; concentrations were measured in water, but data were not presented
Cyprinus carpio, Carp	31 pg/g ⁴		4.8 pg/g ⁴		0.06			[13]	F; Petenwell Reservoir, central Wisconsin; BSAF based on 8% tissue lipid content and 3.1% sediment organic carbon
Salmonids							0.054	[18]	F
Wildlife									
Falco peregrinus, Peregrine falcon			11 ng/g (eggs) (n = 6)	11.4% eggshell thinning				[17]	F; Kola Peninsula, Russia

Species:	Concentration, Units in ¹ :			Toxicity:	Ability to Accumulate ² :			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Haliaeetus leucocephalus, Bald eagle chicks			Powell River site: ~2,800 ng/kg lipid weight basis (yolk sac) Reference site: ~500 ng/kg lipid weight basis (yolk sac)	A hepatic cytochrome P4501A cross-reactive protein (CYP1A) was induced nearly six-fold in chicks from Powell River site compared to the reference (p < 0.05). No significant concentration-related effects were found for morphological, physiological, or histological parameters.				[15]	F; southern coast of British Columbia; eggs were collected from nests and hatched in the lab; ~ indicates value was taken from a figure.	

Species:	Concentrati	ion, Units in¹:		Toxicity:	Ability t	o Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Ardea herodias, Great blue heron chicks			Nicomekl site: 6±2.2 ng/kg (egg) (n = 11) Vancouver site: 57±25.8 ng/kg (egg)	Depression of growth				[16]	L; eggs were collected from three British Columbia colonies with different levels of
			(n = 12)	compared to Nicomekl site. Presence of edema.					contamination and incubated in the laboratory
			Crofton site: 263±69.9 ng/kg (egg) (n = 6)	Depression of growth compared to Nicomekl site. Presence of edema.					

¹ Concentration units based on wet weight unless otherwise noted.
² BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ Not clear from reference if concentration is based on wet or dry weight.

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Chemical Category: POLYCHLORINATED DIBENZO-p-DIOXINS

Chemical Name (Common Synonyms): CASRN: 1746-01-6

2,3,7,8-TETRACHLORODIBENZO-p-DIOXIN

Chemical Characteristics

Solubility in Water: 19.3 ng/L [1] **Half-Life:** 1.1.15 - 1.62 years based on soil

die-away test and lake water and sediment die-away test [2]

Log K_{ov} : 6.53 [3] Log K_{oc} : 6.42 L/kg organic carbon

Human Health

Oral RfD: No data [4] Confidence: —

Critical Effect: —

Oral Slope Factor: 1.5 x 10⁺⁵ per (mg/kg)/day [4] Carcinogenic Classification: B2 [4]

Wildlife

Partitioning Factors: Partitioning factors for 2,3,7,8-TCDD in wildlife were not found in the literature.

Food Chain Multipliers: Limited information was found reporting on specific biomagnification factors of PCDDs and PCDFs through terrestrial wildlife. Due to the toxicity, high K_{ow} values, and highly persistent nature of the PCDDs and PCDFs, they possess a high potential to bioaccumulate and biomagnify through the food web. PCDDs and PCDFs have been identified in fish and wildlife throughout the global aquatic and marine environments [5]. Studies conducted in Lake Ontario indicated that biomagnification of 2,3,7,8-tetrachlorodibenzo-p-dioxin (2,3,7,8-TCDD) appears to be significant between fish and fish-eating birds but not between fish and their food. When calculated for older predaceous fish such as lake-trout-eating young smelt, the biomagnification factor (BMF) can equal 3. The BMF from alewife to herring gulls in Lake Ontario was 32 for 2,3,7,8-TCDD [6].

EPA has developed risk-based concentrations of 2,3,7,8-TCDD in different media that present low and high risk to fish, mammalian, and avian wildlife. These concentrations were developed based on toxic effects of 2,3,7,8-TCDD and its propensity to bioaccumulate in fish, mammals, and birds.

Environmental Concentrations Associated With 2,3,7,8-TCDD Risk to Aquatic Life and Associated Wildlife [7]

	Fish Concentration	Sediment Concentration	Water Conce	ntration (pg/L)					
Organism	(pg/g)	(pg/g dry wt.)	POC=0.2	POC=1.0					
Low Risk									
Fish	50	60	0.6	3.1					
Mammalian Wildlife	0.7	2.5	0.008	0.04					
Avian Wildlife	6	21	0.07	0.35					
	High R	isk to Sensitive Species							
Fish	80	100	1.0	5					
Mammalian Wildlife	7	25	0.08	0.4					
Avian Wildlife	60	210	0.7	3.5					

Note: POC - Particulate organic carbon

Fish lipid of 8% and sediment organic carbon of 3% assumed where needed.

For risk to fish, BSAF of 0.3 used; for risk to wildlife, BSAF of 0.1 used.

Low risk concentrations are derived from no-effects thresholds for reproductive effects (mortality in embryos and young) in sensitive species.

High risk concentrations are derived from TCDD doses expected to cause 50 to 100% mortality in embryos and young of sensitive species.

Aquatic Organisms

Partitioning Factors: Steady-state BSAFs for invertebrates exposed to 2,3,7,8-TCDD in the laboratory ranged from about 0.5 to 0.9 [8]. The BSAF for carp collected from a reservoir in central Wisconsin was 0.27 [9].

Food Chain Multipliers: Biomagnification of 2,3,7,8-TCDD does not appear to be significant between fish and their prey. Limited data for the base of the Lake Ontario lake trout food chain indicated little or no biomagnification between zooplankton and forage fish. BMFs based on fish consuming invertebrate species are probably close to 1.0 because of the 2,3,7,8-TCDD biotansformation by forage fish. BMFs greater than 1.0 may exist between some zooplankton species and their prey due to the lack of 2,3,7,8-TCDD biotansformation in invertebrates [7].

Toxicity/Bioaccumulation Assessment Profile

The polychlorinated dibenzo-*p*-dioxins (PCDDs) and polychlorinated dibenzofurans (PCDFs) each consist of 75 isomers that differ in the number and position of attached chlorine atoms. The PCDDs and PCDFs are polyhalogenated aromatic compounds and exhibit several properties common to this group of compounds. These compounds tend to be highly lipophilic and the degree of lipophilicity is increased with increasing ring chlorination [5]. In general, the PCDDs and PCDFs exhibit relative inertness to acids, bases, oxidation, reduction, and heat, increasing in environmental persistence and chemical stability with increasing chlorination [10,5]. Because of their lipophilic nature, the PCDDs and PCDFs have been detected in fish, wildlife, and human adipose tissue, milk, and serum [5].

Each isomer has its own unique chemical and toxicological properties. The most toxic of the PCDD and PCDF isomers is 2,3,7,8-tetrachlorodibenzo-*p*-dioxin (2,3,7,8-TCDD), one of the 22 possible congeners of tetrachlorodibenzo-*p*-dioxin [10]. Toxicity equivalency factors (TEFs) have been developed by the EPA relating the toxicities of other PCDD and PCDF isomers to that of 2,3,7,8-TCDD [11]. The biochemical mechanisms leading to the toxic response resulting from exposure to PCDDs and PCDFs are not known in detail, but experimental data suggest that an important role in the development of systemic toxicity resulting from exposure to these chemicals is played by an intracellular protein, the Ah receptor. This receptor binds halogenated polycyclic aromatic molecules, including PCDDs and PCDFs. In several mouse strains, the expression of toxicity of 2,3,7,8-TCDD-related compounds, including cleft palate formation, liver damage, effects on body weight gain, thymic involution, and chloracnegenic response, has been correlated with their binding affinity for the Ah receptor, and with their ability to induce several enzyme systems [11].

Toxicity Equivalency Factors (TEF) for PCDD and PCDF Isomers [11]

Isomer	TEF
Total TetraCDD	1
2,3,7,8-TCDD	1
Other TCDDs	0.01
Total PentaCDDs	0.5
2,3,7,8-PentaCDDs	0.5
Other PentaCDDs	0.005
Total HexaCDDs	0.04
2,3,7,8-HexaCDDs	0.04
Other HexaCDDs	0.0004
Total HeptaCDDs	0.001
2,3,7,8-HeptaCDDs	0.001
Other HeptaCDDs	0.00001
Total TetraCDFs	0.1
2,3,7,8-TetraCDF	0.1
Other TetraCDFs	0.001
Total PentaCDFs	0.1
2,3,7,8-PentaCDFs	0.1
Other PentaCDFs	0.001
Total HexaCDFs	0.01
2,3,7,8-HexaCDFs	0.01
Other HexaCDFs	0.0001
Total HeptaCDFs	0.001
2,3,7,8-HeptaCDFs	0.001
Other HeptaCDFs	0.00001

In natural systems, PCDDs and PCDFs are typically associated with sediments, biota, and the organic carbon fraction of ambient waters [7]. Congener-specific analyses have shown that the 2,3,7,8-

substituted PCDDs and PCDFs were the major compounds present in most sample extracts [5]. Results from limited epidemiology studies are consistent with laboratory-derived threshold levels to 2,3,7,8-TCDD impairment of reproduction in avian wildlife. Population declines in herring gulls (*Larus argentatus*) on Lake Ontario during the early 1970s coincided with egg concentrations of 2,3,7,8-TCDD and related chemicals expected to cause reproductive failure based on laboratory experiments (2,3,7,8-TCDD concentrations in excess of 1,000 pg/g). Improvements in herring gull reproduction through the mid-1980s were correlated with declining 2,3,7,8-TCDD concentrations in eggs and lake sediments [7]. Based on limited information on isomer-specific analysis from animals at different trophic levels, it appears that at higher trophic levels, i.e., fish-eating birds and fish, there is a selection of the planar congeners with the 2,3,7,8-substituted positions [12].

PCDDs and PCDFs are accumulated by aquatic organisms through exposure routes that are determined by the habitat and physiology of each species. With log K_{ow}>5, exposure through ingestion of contaminated food becomes an important route for uptake in comparison to respiration of water [7]. The relative contributions of water, sediment, and food to uptake of 2,3,7,8-TCDD by lake trout in Lake Ontario were examined by exposing yearling lake trout to Lake Ontario smelt and sediment from Lake Ontario along with water at a 2,3,7,8-TCDD concentration simulated to be at equilibrium with the sediments. Food ingestion was found to contribute approximately 75 percent of total 2,3,7,8-TCDD [7]. There have been a number of bioconcentration studies of 2,3,7,8-TCDD using model ecosystem and single species exposure. Although there is variation in the actual BCF values, in general, the algae and plants have the lowest BCF values, on the order of a few thousand. A log BCF value of 4.38 has been reported for the snail *Physa* sp. Crustacea and insect larvae appear to have the next highest BCF values, followed by several species of fish, with the highest log BCF value of 4.79 [12].

Exposure of juvenile rainbow trout to 2,3,7,8-TCDD and -TCDF in water for 28 days resulted in adverse effects on survival, growth, and behavior at extremely low concentrations. A no-observed-effects concentration (NOEC) for 2,3,7,8-TCDD could not be determined because the exposure to the lowest dose of 0.038 ng/l resulted in significant mortality [13]. A number of biological effects have been reported in fish following exposure to 2,3,7,8-TCDD including enzyme induction, immunological effects, wasting syndrome, dermatological effects, hepatic effects, hematological effects, developmental effects, and cardiovascular effects [12].

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Plants									
Oedogonium cardiacum, Green algae			1.34 mg/kg (whole body) ⁴	Growth, NOED				[35]	L; no effect on growth
Lemna minor, Duckweed			0.00614 mg/kg (whole body) ⁴	Growth, NOED				[35]	L; no observed effect
Invertebrates									
Nereis virens, Sandworm	656±97 pg/g dw; (n = 6)		422±103 pg/g dw (whole body)				~0.5	[8,14]	L; 180-day exposure; sediment TOC was 57 mg/kg; ~ indicates approximate value, as numbers were estimated from bar graphs.
<i>Physa</i> sp., Snail			0.364 mg/kg (whole body) ⁴	Mortality, NOED				[35]	L; no effect on survival

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability	to Accumul	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Macoma nasuta, Clam	656±97 pg/g dw; (n = 6)		142 ± 20 pg/g dw				~0.9	[8,14]	L; 120-day exposure; sediment TOC was 57 mg/kg; ~ indicates approximate value, as numbers were estimated from bar graphs.
Daphnia magna, Cladaceran			2.08 mg/kg (whole body) ⁴	Mortality, NOED				[35]	L; no effect on survival
Palaemonetes pugio, Grass shrimp	656±97 pg/g dw; (n = 6)		138 ± 20 pg/g dw				~0.7	[8,14]	L; 28-day exposure; sediment TOC was 57 mg/kg ~ indicates approximate value, as numbers were estimated from bar graphs.
Pacifastacus leniusculus, Crayfish			0.003 mg/kg (whole body) ⁴	Mortality, ED25				[31]	L; 25% mortality after 40 days
٠			0.03 mg/kg (whole body) ⁴	Mortality, ED50				[31]	L; lethargy, 50% to 66% increase in mortality

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability t	o Accumul	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.003 mg/kg (whole body) ⁴	Behavior, LOED				[31]	L; lack of avoidance response
			0.003 mg/kg (whole body) ⁴	Physiological, LOED				[31]	L; significant induction of cytochrome P450
			0.003 mg/kg (whole body) ⁴	Physiological, LOED				[31]	L; significant induction of liver enzymes (cytochrome P450)
			0.1 mg/kg (whole body) ⁴	Morphology, NOED				[31]	L; no significant pathology at highest dose
			0.0003 mg/kg (whole body) ⁴	Mortality, NOED				[31]	L; no effect on mortality
			0.0003 mg/kg (whole body) ⁴	Physiological, NOED				[31]	L; no significant induction of liver enzymes (cytochrome P450)
Callinectes sapidus, Blue crab	32.2 ppt^5 (TOC = 3.2%) 52.8 ppt^5 (TOC = 3.9%)	,	8.2 ppt ⁵ (hepatopancreas) (% lipid = 7.6)			-0.72	0.089	[15]	F; northeastern Florida; bleach- kraft paper mill receiving stream; BAF and BSAF calculated using mean of two sediment concentrations.

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumula	te ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Chironomus tentans, Midge			0.47 mg/kg (whole body)	Growth, NOED				[44]	L; concentrations are lipid
Fishes									
Oncorhynchus mykiss (Salmo gairdneri),		water exposure 0.038 ng/L	1.0 μg/kg ⁵	28-day LOEC (survival, growth)	4.41			[13]	L
Rainbow trout		water exposure 0.382 ng/L	10.95 ng/g ⁵ (whole body)		4.46				L; 6-hour exposure period
Oncorhynchus mykiss, Rainbow trout			0.00388 mg/kg (extractable lipid) ⁴	Growth, LOED				[32]	L; reduced growth, exposed fish weighed 50 g vs. 130 g for controls
			0.00371 mg/kg (liver) ⁴	Growth, LOED				[32]	L; reduced growth, exposed fish weighed 50 g vs. 130 g for controls
			0.00026 mg/kg (muscle) ⁴	Growth, LOED				[32]	L; reduced growth, exposed fish weighed 50 g vs. 130 g for controls

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumula	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.00065 mg/kg (whole body) ⁴	Growth, LOED				[32]	L; reduced growth, exposed fish weighed 50 g vs. 130 g for controls
			0.00388 mg/kg (etractable lipid) ⁴	Morphology, LOED				[32]	L; livers enlarged to nearly twice the size of control fish livers, fin rot
			0.00371 mg/kg (liver) ⁴	Morphology, LOED				[32]	L; livers enlarged to nearly twice the size of control fish livers, fin rot
			0.00026 mg/kg (muscle) ⁴	Morphology, LOED				[32]	L; livers enlarged to nearly twice the size of control fish livers, fin rot
			0.00065 mg/kg (whole body) ⁴	Morphology, LOED				[32]	L; livers enlarged to nearly twice the size of control fish livers, fin rot
			0.00388 mg/kg (extractable lipid) ⁴	Mortality, LOED				[32]	L; lethal to 7 of 90 fish over 139 days
			0.00371 mg/kg (liver) ⁴	Mortality, LOED				[32]	L; lethal to 7 of 90 fish over 139 days
			0.00026 mg/kg $(\text{muscle})^4$	Mortality, LOED				[32]	L; lethal to 7 of 90 fish over 139 days

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability	to Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.00065 mg/kg (whole body) ⁴	Mortality, LOED				[32]	L; lethal to 7 of 90 fish over 139 days
Oncorhynchus mykiss,			0.01 mg/kg (whole body) ⁴	Mortality, ED50				[36]	L; 80-day LD50 for mortality
Rainbow trout			0.001 mg/kg (whole body) ⁴	Growth, LOED				[36]	L; reduction in body weight
			0.025 mg/kg (whole body) ⁴	Morphology, LOED				[36]	L; fin necrosis, hyperpigmentation
Oncorhynchus mykiss, Rainbow trout			0.000315 mg/kg (carcass) ⁴	Growth, NOED				[38]	L; no effect on growth
Rambow trout			0.000102 mg/kg (gastrointestinal tract) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.000244 mg/kg (gill) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.00007 mg/kg (heart) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.000092 mg/kg (kidney) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.000072 mg/kg (liver) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.000355 mg/kg (pyloric caeca) ⁴	Growth, NOED				[38]	L; no effect on growth

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumula	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000029 mg/kg (skeletal muscle) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.000201 mg/kg (skin) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.000085 mg/kg (spleen) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.00327 mg/kg (visceral fat) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.00025 mg/kg (whole body) ⁴	Growth, NOED				[38]	L; no effect on growth
			0.000315 mg/kg (carcass) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000102 mg/kg (gastrointestinal tract) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000244 mg/kg (gill) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.00007 mg/kg (heart) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000092 mg/kg (kidney) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000072 mg/kg (liver) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000355 mg/kg (pyloric caeca) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000029 mg/kg (skeletal muscle) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000201 mg/kg (skin) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000085 mg/kg (spleen) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.00327 mg/kg (visceral fat) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.00025 mg/kg (whole body) ⁴	Morphology, NOED				[38]	L; no effect on fin necrosis or hemorrhage
			0.000315 mg/kg (carcass) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.000102 mg/kg (gastrointestinal tract) ⁴	Mortality, NOED				[38]	L; no effect on mortality

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000244 mg/kg (gill) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.00007 mg/kg (heart) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.000092 mg/kg (kidney) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.000072 mg/kg (liver) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.000355 mg/kg (pyloric caeca) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.000029 mg/kg (skeletal muscle) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.000201 mg/kg (skin) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.000085 mg/kg (spleen) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.00327 mg/kg (visceral fat) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.00025 mg/kg (whole body) ⁴	Mortality, NOED				[38]	L; no effect on mortality
			0.00452 mg/kg (whole body) ⁴	Survival, ED50				[13]	L; exposure concentration is the mean of measured TCDD concentration

Species:	Concentrat	Concentration, Units in ¹ :			Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Oncorhynchus mykiss, Rainbow trout			0.0000047 mg/kg (liver) ⁴	Biochemical, LOED				[40]	L; significant increase in liver ethoxyresorufin O- deethylase (EROD
			0.000038 mg/kg (liver) ⁴	Biochemical, LOED				[40]	L; significant increase in liver ethoxyresorufin Odeethylase (EROD)
			0.000016 mg/kg (liver) ⁴	Biochemical, LOED				[40]	L; significant increase in liver ethoxyresorufin O- deethylase (EROD
Oncorhynchus mykiss, Rainbow trout			0.000439 mg/kg (whole body) ⁴	Mortality, ED50				[42]	L; mortality from fertilization to swim-up; exposure dose calculated from text; residue measured in egg at 5-days post exposure; dosed fo 48 hours and endpoint measured after approximately 24 days

Species:	Concentrat	ion, Units in	1:	Toxicity:	Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000421 mg/kg (whole body) ⁴	Mortality, ED50				[42]	L; mortality from fertilization to swim-up; liposome used to carry dose; 93% of dose retained in egg and assumed to be in swim-up fry, flow rate = 8-12
			0.000279 mg/kg (whole body) ⁴	Mortality, LOED				[42]	L; significant increase in mortality from hatch to swim-up at lowest exposure concentration tested; exposure dose calculated from text; residue measured in egg at 5-days post exposure; dosed for 48 hours and endpoint measured after approximately 24 days

Species:	Concentrat	tion, Units in¹:		Toxicity:	Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000437 mg/kg (whole body) ⁴	Mortality, LOED				[42]	L; significant increase in mortality from hatch to swim-up; liposome used to carry dose; 93% of dose retained in egg and assumed to be in swim-up fry, flow rate = 8-12
			0.000291 mg/kg (whole body) ⁴	Mortality, NOED				[42]	L; no significant increase in mortality from hatch to swim-up; liposome used to carry dose; 93% of dose retained in egg and assumed to be in swim-up fry, flow rate = 8 to 12
Oncorhynchus mykiss, Rainbow trout			0.00017 mg/kg (whole body) ⁴	Mortality, ED50				[43]	L; estimated LD50s for 6 strains of rainbow trout, orig_con_wet ranged from 170 to 488; used low value; exposure concentration = 170 to 488

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability 1	to Accumula	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Salmo trutta, Brown trout			5.2 pg/g ⁵ (fillet)			4.25-4.45		[22]	F; locations throughout Maine; a range of mean BAFs is presented; the values are means for locations throughout Maine, and the range is for BAFs calculated using river concentrations from years prior to the sampling date to account for declines in paper mill discharges
Salvelinus fontinalis, Brook trout			0.0006 mg/kg (whole body) ⁴	Physiological				[33]	L; induction of hepatic EROD

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumula	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Salvelinus fontinalis, Brook trout			0.0012 mg/kg (whole body) ⁴	Growth, NOED				[41]	L; no significant growth effect at highest target body burden; TCDD-spiked diet to produce desired body burden; abstract with minimal information
			0.0012 mg/kg (whole body) ⁴	Mortality, NOED				[41]	L; no significant mortality at highest target body burden; TCDD-spiked diet to produce desired body burden; abstract with minimal information
			0.0012 mg/kg (whole body) ⁴	Reproduction, LOED				[41]	L; significant delay in spawning; TCDD-spiked diet to produce desired body burden; abstract with minimal information

Species:	Concentrat	tion, Units in¹:		Toxicity:	Ability t	o Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.0006 mg/kg (whole body) ⁴	Reproduction, NOED				[41]	L; no delay in spawning; TCDD- spiked diet to produce desired body burden; abstract with minimal information
Salvelinus fontinalis, Brook trout			0.0002 mg/kg (whole body) ⁴	Mortality, ED50				[43]	L; estimated LD50
<i>Amia calva</i> , Bowfin			11.2 ppt ⁵ (liver) (n = 1)			-0.59	0.180	[15]	F; northeastern Florida; bleached-
			18.6 ppt ⁵ (liver) (n = 1)			-0.36	0.255		kraft paper mill receiving stream; BAF and BSAF calculated using
			46.1 ppt ⁵ (ovary) (n=1)			0.03	0.281		mean of two sediment concentrations. BAF = (pg TCDD/g tissue) ÷ (pg TCDD / g sediment); BSAF = (pg TCDD/g lipid) ÷ (pg TCDD / g TCDD / g TCDD / g TCDD / g TCDD /

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumula	nte ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Oncorhynchus kisutch, Coho salmon			0.000478 mg/kg (whole body) ⁴	Growth, LOED				[39]	L; reduced growth
			0.000478 mg/kg (whole body) ⁴	Mortality, LOED				[39]	L; reduced survival
			0.00217 mg/kg (whole body) ⁴	Growth, NA				[39]	L; reduced growth
			0.00217 mg/kg (whole body) ⁴	Mortality, NA				[39]	L; reduced survival
			0.000125 mg/kg (whole body) ⁴	Behavior, NOED				[39]	L; no effect on food consumption or feeding
			0.000125 mg/kg (whole body) ⁴	Growth, NOED				[39]	L; no effect on growth
Carp, (scientific name unknown)		water exposure 62 pg/L	2.2 μg/kg ⁵	Death (71 days)				[16]	L
Salvelinus namaycush, Lake trout, (early life stage)		water exposure 20 ng/L	$0.055 \mu\mathrm{g/kg^5}(\mathrm{egg})$	48-hour LOEC (mortality)				[21]	L

Species:	Concentrat	ion, Units in¹:	Toxicity:	Ability t	o Accumu	late ² :	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
		water exposure 10 ng/L	$0.034 \mu\text{g/kg}^5 (\text{egg})$	48-hour NOEC (mortality)				[21]	L
		water exposure 62 ng/L	$0.226 \mu\mathrm{g/kg^5}$ (egg)	48-hour LOEC (hatchability)				[21]	L
Salvelinus namaycush, Lake trout			0.000065 mg/kg (whole body) ⁴	Mortality, ED50				[15]	L; lethal to 50% of sac fry
			0.000055 mg/kg (whole body) ⁴	Mortality, LOED				[15]	L; lowest statistically significant increase in mortality of sac fry
			0.000226 mg/kg (whole body) ⁴	Reproduction, L				[15]	L; reduced hatchability of eggs
			0.000035 mg/kg (whole body) ⁴	Mortality, NOED				[15]	L; no effect on mortality of sac fry
Salvelinus namaycush, Lake trout			0.000044 mg/kg (whole body) ⁴	Mortality, ED50				[34]	L; LD50 for sac fry mortality

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Salvelinus namaycush, Lake trout			0.000065 mg/kg (whole body) ⁴	Mortality, ED50				[42]	L; mortality from fertilization to swim-up; exposure dose calculated from text; residue measured in egg at 5-days post exposure; dosed for 48 hours and endpoint measured after approximately 24 days
			0.000055 mg/kg (whole body) ⁴	Mortality, LOED				[42]	L; significant increase in mortality from hatch to swim-up; exposure dose calculated from text; residue measured in egg at 5-days post exposure; dosed for 48 hours and endpoint measured after approximately 24 days

Species:	Concentrat	ion, Units in ¹	:	Toxicity:	Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000058 mg/kg (whole body) ⁴	Mortality, LOED				[42]	L; significant increase in mortality from hatch to swim-up; high control mortality; liposomused to carry dose; 93% of dose retained in egg and assumed to be in swim-up fry, flow rate = 8-12
			0.000034 mg/kg (whole body) ⁴	Mortality, NOED				[42]	L; no significant increase in mortality from hatch to swim-up; exposure dose calculated from text; residue measured in egg at 5-days post exposure; dosed for 48 hours and endpoint measured after approximately 24 days

Species:	Concentration, Units in¹:			Toxicity:	Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000044 mg/kg (whole body) ⁴	Mortality, NOED				[42]	L; no significant increase in mortality from hatch to swim-up; high control mortality; liposome used to carry dose; 93% of dose retained in egg and assumed to be in swim-up fry, flow rate = 8 to 12
Salvelinus namaycush, Lake trout			0.000065 mg/kg (whole body) ⁴	Mortality, ED50				[43]	L; estimated LD50
Carassius auratus, Goldfish			0.58-0.63 ng/g ⁵ (whole body)		4.39			[18]	L; fish were exposed for 120 hr; exposure water contained fly ash extract; concentrations were measured in water, but data were not presented

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability t	o Accumul	late²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Cyprinus carpio, Carp	170 pg/g⁵		120 pg/g ⁵				0.27	[9]	F; Petenwell Reservoir, central Wisconsin; BSAF based on 8% tissue lipid content and 3.1% sediment organic carbon
Cyprinus carpio, Carp			0.0022 mg/kg (whole body) ⁴	Behavior, LOED				[15]	L; difficulty swimming
			0.0022 mg/kg (whole body) 4	Cellular, LOED				[15]	L; edema, body wall ulcers
			0.0022 mg/kg (whole body) ⁴	Morphology, LOED				[15]	L; fin erosion, hemorrhage, morphologically resembling Blue Sac Disease
Cyprinus carpio, Carp			0.003 mg/kg (whole body) ⁴	Mortality, ED50				[36]	L; 80-day LD50 for mortality
Cyprinus carpio, Carp			0.0022 mg/kg (whole body) ⁴	Mortality, LOED				[15]	L; increased mortality

Species:	Concentrat	tion, Units in¹:		Toxicity:	Ability to	o Accumul	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Danio rerio, Zebrafish			2.16 ng/g (egg)	ED50 (pericardial edema)				[23]	L; newly fertilized eggs were exposed for 1 hr to water
			2.43 ng/g (egg)	ED50 (yolk sac edema)					containing graded concentrations of
			2.45 ng/g (egg)	LD50					TCDD
Bracydanio rerio, Zebrafish			8.3 μg/kg ⁵	LOEC (reproduction)				[24]	L; food exposure
			8.3 μg/kg ⁵	LOEC (oogenesis)				[24]	L; food exposure
Pimephales promelas, Fathead minnow			17-2,042 μg/kg ⁵	LD100				[17]	L; food exposure

Species:	Concentration, Units in ¹ :			Toxicity:	Ability	Ability to Accumulate ² :			Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Catostomus commerson, White sucker			9.6 pg/g ⁵ (whole body)			4.89-5.0	3	[22]	F; locations throughout Maine; a range of mean BAFs is presented; the values are means for locations throughout Maine, and the range is for BAFs calculated using river concentrations from years prior to the sampling date to account for declines in paper mill discharges	

Species:	Concentration	ncentration, Units in¹: Toxicity: Ability to Accumulate²:		ate ² :	Source:				
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Ictalurus nebulosus, Brown bullhead catfish	, 32.2 -52.8 ppt ⁵ (TOC = 3.2- 3.9%)		1.8 ppt ⁵ (liver) (% lipid = 3.5)			-1.40	0.043	[15]	F; northeastern Florida; bleached- kraft paper mill receiving stream; BAF and BSAF calculated using mean of two sediment concen- trations. BAF = (pg TCDD/g tissue) ÷ (pg TCDD/ g sediment); BSAF = (pg TCDD/g lipid) ÷ (pg TCDD / g TCDD / g TCDD / g TCDD / g TCDD /
catrisn			2.6 ppt ⁵ (liver) (% lipid = 2.9)			-1.22	0.074		
			2.8 ppt ⁵ (liver) (% lipid = 3.2)			-1.15	0.073		
Ictalurus melas, Black bullhead			0.005 mg/kg (whole body) ⁴	Mortality, ED50				[36]	L; 80 day LD50 for mortality
			0.025 mg/kg (whole body) ⁴	Morphology, LOED				[36]	L; fin necrosis, hyperpigmentation

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability 1	to Accumu	late²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Ictalurus punctatus, Channel catfish			0.0044 mg/kg (whole body) ⁴	Mortality, ED10				[45]	L; radiolabelled compounds in sediment, compound leached into water for exposure; all fish died between days 14 and 15; body residues from dead fish
Gambusia affinis, Mosquito fish			0.0072 mg/kg (whole body) ⁴	Mortality, ED10				[45]	L; radiolabelled compounds in sediment, compound leached into water for exposure; all fish died between days 14 and 15; body residues from dead fish
Oryzias latipes, Japanese medaka		Water exposure 2.2 ng/L	0.24 μg/kg ⁵ (embryo)	Lesions in embryos				[19]	L

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	Accumul	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Oryzias latipes, Japanese medaka (juveniles)		water exposure 101 ± 26 pg/L (n = 23)	2,408 ±241 pg/g	Obvious signs of TCDD toxicity such as generalized edema, fin erosion, and discoloration in many of the exposed fish	4.38 non- steady state 5.71 predicted steady state			[20]	L; 12-day exposure period; lipid content 7.5%
Oryzias latipes, Japanese medaka			0.24 mg/kg (whole body) ⁴	Lesions, ED50				[19]	L; ten replicates per treatment
			0.3 mg/kg (whole body) ⁴	Lesions, LOED				[19]	L; 50% of embryos with lesions but no statistical significance analyzed; ten replicates per treatment
			0.1 mg/kg (whole body) ⁴	Lesions, NOED				[19]	L; no significant incidence of lesions at lowest doseage tested; 10 replicates per treatment, resd_conc_wet value > 0.1

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	Ability to Accumulate ² :			Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Morone americana, White perch			1.2 pg/g ⁵			3.48-3.88		[22]	F; locations throughout Maine; a range of mean BAFs is presented; the values are means for locations throughout Maine, and the range is for BAFs calculated using river concentrations from years prior to the sampling date to account for declines in paper mill discharges	
Lepomis macrochirus,			0.016 mg/kg (whole body) ⁴	Mortality, ED50				[36]	L; 80-day LD50 for mortality	
Bluegill			0.005 mg/kg (whole body) ⁴	Growth, LOED				[36]	L; reduction in body weight	
			0.025 mg/kg (whole body) ⁴	Morphology, LOED				[36]	L; fin necrosis, hyperpigmentation	

Species:	Concentration, Units in ¹ :			Toxicity:	Ability to Accumulate ² :			Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Microperus dolomieu, Smallmouth bass			3.4 pg/g ⁵ (fillet)			4.06-4.39		[22]	F; locations throughout Maine; a range of mean BAFs is presented; the values are means for locations throughout Maine, and the range is for BAFs calculated using river concentrations from years prior to the sampling date to account for declines in paper mill discharges

Species:	Concentration	on, Units in¹:		Toxicity:	Ability to Accumulate ² :			Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Micropterus samoides, Largemouth bass	32.2 ppt ⁵ (TOC=3.2%)		1.8 ppt ⁵ (liver) (% lipid =3.9)			-1.40	0.038	[15]	F; northeastern Florida; bleached kraft paper mill
Zargemouti ouss	52.8 ppt ⁵ (TOC=3.9%))	2.9 ppt ⁵ (liver) (% lipid =2.4)			-1.15	0.100		receiving stream; BAF and BSAF calculated using
			8.8 ppt ⁵ (ovary) (% lipid =7.6)			-0.68	0.096		mean of two sediment concentrations. BAF = (pg TCDD / g tissue) ÷ (pg TCDD/ g sediment); BSAF = (pg TCDD/g lipid) ÷ (pg TCDD/ g TOC).
Micropterus salmoides, Largemouth bass			0.011 mg/kg (whole body) ⁴	Mortality, ED50				[36]	L; 80-day LD50 For Mortality
			0.025 mg/kg (whole body) ⁴	Morphology, LOED				[36]	L; Fin Necrosis, Hyperpigment- ation
Perca flavescens, Yellow perch			0.003 mg/kg (whole body) ⁴	Mortality, ED50				[36]	L; 80-day LD50 for mortality
			0.005 mg/kg (whole body) ⁴	Growth, LOED				[36]	L; reduction in body weight

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability to	o Accumula	te ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.025 mg/kg (whole body) ⁴	Morphology, LOED				[36]	L; fin necrosis, hyperpigmentation
			0.000129 mg/kg (carcass) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000148 mg/kg (gastrointestinal tract) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000155 mg/kg (gill) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000077 mg/kg (heart) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000119 mg/kg (kidney) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000466 mg/kg (liver) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000143 mg/kg (pyloric caeca) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000009 mg/kg (skeletal muscle) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000041 mg/kg (skin) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000166 mg/kg (spleen) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.00277 mg/kg (visceral fat) ⁴	Growth, NOED				[37]	L; no effect on growth

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability t	o Accumul	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000143 mg/kg (whole body) ⁴	Growth, NOED				[37]	L; no effect on growth
			0.000129 mg/kg (carcass) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000148 mg/kg (gastrointestinal tract) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000155 mg/kg (gill) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000077 mg/kg (heart) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000119 mg/kg (kidney) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000466 mg/kg (liver) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000143 mg/kg (pyloric caeca) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000009 mg/kg (skeletal muscle) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability (to Accumula	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000041 mg/kg (skin) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000166 mg/kg (spleen) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.00277 mg/kg (visceral fat) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000143 mg/kg (whole body) ⁴	Morphology, NOED				[37]	L; no effect on fin necrosis or hemorrhage
			0.000129 mg/kg (carcass) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000148 mg/kg (gastrointestinal tract) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000155 mg/kg (gill) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000077 mg/kg (heart) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000119 mg/kg (kidney) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000466 mg/kg (liver) ⁴	Mortality, NOED				[37]	L; no effect on mortality

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability 1	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.000143 mg/kg (pyoric ceca) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000009 mg/kg (skeletal muscle) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000041 mg/kg (skin) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000166 mg/kg (spleen) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.00277 mg/kg (visceral fat) ⁴	Mortality, NOED				[37]	L; no effect on mortality
			0.000143 mg/kg (whole body) ⁴	Mortality, NOED				[37]	L., no effect on mortality
Salmonids							0.059	[46]	F

Species:	Concentrat	tion, Units in	¹ :	Toxicity:	Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Wildlife									
Aix sponsa, Wood duck			pg/g (eggs):	% eggs hatched:				[29]	F; central Arkansas; egg
			Site 1 geometric mean: 36 (1.6 to 482)	47% (9.7 SE)					TEFs and hatching success and
			Site 2 geometric mean: 14 (0.8-74)	62% (10.1 SE)					duckling production were negatively correlated; clutch
			Site 3 geometric mean: 4.2 (<1 to 19)	79% (3.8 SE)					size was similar among wetland Sites 1-3 which
			Site 4 geometric mean: 0.01 (<1 to 0.5)	93% (3.4 SE)					were 9, 17, and 58 km downstream from point source of contamination. respectively, and Site 4 which was 111 km away on a separate drainage; duckling abnormalities were also noted
Aix sponsa, Wood duck									Threshold range of reduced productivity was > 20-50 ppt TEF.

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability	to Accumu	ılate²:	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Falco peregrinus, Peregrine falcon			11 ng/g (eggs) (n= 6)	11.4% eggshell thinning				[26]	F; Kola Peninsula, Russia	
Haliaeetus leucocephalus, Bald eagle chicks			Powell River site: 2,200 ng/kg lipid weight basis (yolk sac) Reference site: 300 ng/kg lipid weight basis (yolk sac)	A nearly 6-fold greater incidence of an hepatic cytochrome P4501A cross-reactive protein was induced in chicks from Powell River site as compared to the reference (p < 0.05). No significant concentration-related effects were found for morphological, physiological, or histological parameters.				[25]	F; southern coast of British Columbia; eggs were collected from nests and hatched in the lab; ~ indicates value was taken from a figure.	

Species:	Concentration, Units in ¹ :			Toxicity:	Ability to	o Accumul	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Sterna forsteri, Forster's tern			Lake Poygan site: 8.0 pg/g; (egg) (n = 6) Green Bay site: 37.3 pg/g; (egg) (n = 6)	Birds from Green Bay had increased incubation period, reduced hatchability, lower body weight, increased liver to body weight ratio, and occurrence of edema when compared to birds from Lake Poygan. There was a significantly higher incidence of congenital abnormalities in dead embryos and chicks from Green Bay.				[27]	F; Green Bay, Lake Michigan, and Lake Poygan, Wisconsin

Species:	Concentrati	ion, Units in¹:		Toxicity:	Ability to	o Accumula	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Ardea herodias, Great blue heron chicks			Nicomekl site: 10±0.9 ng/kg; (egg) (n = 11)					[28]	L; eggs were collected from three British Columbia colonies
			Vancouver site: 135±49.6 ng/kg (egg) (n = 12)	Depression of growth compared to Nicomekl site. Presence of edema.					with different levels of contamination and incubated in the laboratory
			Crofton site: 2 11±33.7 ng/kg (egg) (n = 6)	Depression of growth compared to Nicomekl site. Presence of edema.					

Species:	Concentration	on, Units in¹:		Toxicity:	Ability 1	to Accumul	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Mustela vison, Mink	Diet: 2 pg/g ⁵		21 pg/g ⁵ (liver)	LOAEL; reduced kit body weights followed by reduced survival		log BMF= 1.05		[30]	L; BMF = biomagnification factor = v_1/v_d , v_1 = lipid-normalized concentration in
	Diet: 3 pg/g ⁵		34 pg/g ⁵ (liver)	Reduced kit body weights followed by reduced survival		log BMF = 1.06			tissue; $v_d = lipid$ - normalized dietary concentration
	Diet: 7 pg/g ⁵		50 pg/g ⁵ (liver)	Significant decrease in number of live kits whelped per female		log BMF = 1.04			

¹ Concentration units based on wet weight unless otherwise noted.

² BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

 $^{^{3}}$ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ This entry was excerpted directly from the Environmental Residue-Effects Database (ERED, www.wes.army.mil/el/ered, U.S. Army Corps of Engineers and U.S. Environmental Protection Agency). The original publication was not reviewed and the reader is strongly urged to consult the publication to confirm the information presented here.

⁵ Not clear from reference if concentration is based on wet or dry weight.

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CASRN: 72-54-8

Chemical Category: PESTICIDE (ORGANOCHLORINE)

Chemical Name (Common Synonyms):

1,1'-(2,2-DICHLOROETHYLIDENE)BIS(4-CHLOROBENZENE),

p,p'-DICHLORODIPHENYLDICHLOROETHANE

4.4'-DICHLORODIPHENYLDICHLOROETHANE

Chemical Characteristics

Solubility in Water: 0.16 mg/L at 24°C [1] **Half-Life:** 2.0-15.6 years based on

biodegradation of DDD in aerobic soils under field conditions [2]

Log K_{ow}: 6.10 [3] **Log K_{oc}:** 6.0 L/kg organic carbon

Human Health

Oral RfD: Not available [4] **Confidence:** —

Critical Effect: Lung tumors in male and female mice, liver tumors in male mice, thyroid tumors in male rats

Oral Slope Factor: 2.4 x 10⁻¹ per (mg/kg)/day [4] Carcinogenic Classification: B2 [4]

Wildlife

Partitioning Factors: Partitioning factors for DDD in wildlife were not calculated in the studies reviewed. However, based on the data presented in one study reviewed, log BCFs for birds from the lower Detroit River ranged from 4.97 to 5.22. Concentrations of DDD in birds were 3.5 to 6.1 times higher than those in sediment.

Food Chain Multipliers: Biomagnification factors of 3.2 and 85 were determined for DDT and DDE, respectively, from alewife to herring gulls in Lake Ontario [5]. A study of arctic marine food chains measured biomagnification factors for DDE that ranged from 17.6 to 62.2 for fish to seal, 0.3 to 0.7 for seal to bear, and 10.7 for fish to bear [6].

Aquatic Organisms

Partitioning Factors: Partitioning factors for DDD in aquatic organisms were not calculated in the studies reviewed. However, the data from one study reviewed showed BCFs of 17,600 for oligochaetes and 565,000 for carp. Ratios of DDD in tissue to sediment were 0.65 for oligochaetes and 21 for carp. BSAFs for clams ranged from 0.120 to 2.745 [22,25]. BSAFs for fish ranged from 0.079 to 2.379 [21,23,24,25].

Food Chain Multipliers: Food chain multipliers (FCMs) for trophic level 3 aquatic organisms were 18.5 (all benthic food web), 1.6 (all pelagic food web), and 11.3 (benthic and pelagic food web). FCMs for trophic level 4 aquatic organisms were 37.4 (all benthic food web), 3.1 (all pelagic food web), and 17.8 (benthic and pelagic food web) [28].

Toxicity/Bioaccumulation Assessment Profile

DDT is very persistent in the environment due to its low vapor pressure, high fat solubility, and resistance to degradation and photooxidation. DDT is degraded to DDE under aerobic conditions and to DDD in anoxic systems [7]. These metabolites, DDD and DDE, are similar to DDT in both their stability and toxicity. Chronic effects of DDT and its metabolites on ecological receptors include changes in enzyme production, hormonal balance, and calcium metabolism, which may cause changes in behavior and reproduction. The high octanol-water partition coefficient of DDT indicates that it is easily accumulated in tissues of aquatic organisms. Laboratory studies have shown that these compounds are readily bioconcentrated in aquatic organisms, with reported log BCFs for DDT ranging from 3.08 to 7.65 and for DDE ranging from 4.80 to 5.26 [8].

Invertebrate species are generally more susceptible than fish species to effects associated with exposure to DDT in the water column [8]. In general, the low solubility of DDT and its metabolites in water suggests that water column exposures are likely to be lower than exposures from ingestion of food or sediment. Sediments contaminated with pesticides, including DDT, have been shown to affect benthic communities at low concentrations. Results of laboratory and field investigations suggest that chronic effects generally occur at total DDT concentrations in sediment exceeding 2 μ g/kg [9]. Equilibrium partitioning methods predict that chronic effects occur at DDT concentrations in sediment of 0.6 to 1.7 μ g/kg [10].

For fish, the primary route of uptake is via prey items, but both DDT and its metabolites can be accumulated through the skin or gills upon exposure to water. Short-term exposure to DDT concentrations of less than $1 \mu g/L$ have been reported to elicit toxic responses in both freshwater and marine fish [8]. DDT may also be transfered to embryos from contaminated adults. DDT concentrations of 1.1 to 2.4 mg/kg in fish embryos have been associated with fry mortality [11,12].

Eggshell thinning, embryo mortality, and decreased hatchling survival have been linked to chronic exposure to DDT and its metabolites in the diet of birds. Of the three compounds, evidence strongly indicates that DDE is responsible for most reproductive toxicity in avian species [13]. Measurements of residues in eggs of birds are a reliable indicator of adverse effects. There is a large amount of variability in sensitivity to DDT and its metabolites among bird species, with waterfowl and raptor species showing the greatest sensitivities. Studies have shown the brown pelican to be most susceptible to adverse effects, with eggshell thinning and depressed productivity occurring at $3.0 \mu g/g$ of DDE in the egg and total reproductive failure when residues exceed $3.7 \mu g/g$ [13].

Species:	Concentration	on, Units in¹:		Toxicity:	Ability t	o Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Invertebrates									
Tubifex sp., Oligochaetes	0.023 mg/kg n = 1	water = 0.85 ng/L n = 1	0.015 mg/kg n = 1					[14]	F; lower Detroit River
<i>Macomona liliana</i> , Mollusk	66.7 μg/kg OC		76.3 μg/kg lipid				1.144	[22]	F; %lipid = 2.95; %sed OC = 0.30
	1,096.0 μg/kg OC		765.2 μg/kg lipid				0.698	[22]	F; %lipid = 2.33; %sed OC = 0.73
	286.4 μg/kg OC		75.1 µg/kg lipid				0.262	[22]	F; %lipid = 2.57; %sed OC = 0.22
	20.0 μg/kg OC		54.9 µg/kg lipid				2.745	[22]	F; %lipid = 2.04; %sed OC = 0.25
	25.0 μg/kg OC		22.4 μg/kg lipid				0.894	[22]	F; %lipid = 3.13; %sed OC = 0.48
Austrovnus stutchburyi, Mollusk	66.7 μg/kg OC		42.4 μg/kg lipid				0.635	[22]	F; %lipid = 5.62; %sed OC = 0.30
	286.4 μg/kg OC		34.4 μg/kg lipid				0.120	[22]	F; %lipid = 4.85; %sed OC = 0.22
	20 μg/kg OC		27.7 μg/kg lipid				1.383	[22]	F; %lipid = 3.87; %sed OC = 0.25
	25 μg/kg OC		25.1 μg/kg lipid				1.002	[22]	F; %lipid = 4.27; %sed OC = 0.48

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	Ability to Accumulate ² :			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³		
Corbicula fluminea, Asian clam	58.8 μg/kg OC		82 μg/kg lipid				1.394	[25]	F; %lipid = 0.61; %sed OC = 1.19		
	159.7 μg/kg OC		82 μg/kg lipid				0.513	[25]	F; %lipid = 0.61; %sed OC = 1.19		
Fish											
Anguilla anguilla, Eel	126 μg/kg OC		10 μg/kg lipid				0.079	[26]	F; %lipid = 13; %sed OC = 32		
Corogonus autumnalis, Omul (endemic whitefish)		particulate: 1.0 pg/L ± 1.0 n = 7	0.0086-0.15 mg/kg lipid (whole body) n = 1								
		dissolved: 17 pg/L ±7.3 n = 7									
Oncorhynchus, Salmo, Salvelinus sp., Salmonids	2,667 μg/kg OC		754.5 µg/kg lipid				0.283	[24]	F; %lipid = 11; %sed OC = 2.7		
		0.000093 μg/L	83 µg/kg			5.93		[24]	F; %lipid = 11		
Salvelinus fontinalis, Brook trout			4.79 mg/kg (whole body) ⁴	Behavior, NOED				[18]	L; temperature selection after 24 h exposure to chemical		

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability 1	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Salvelinus namaycush, Lake trout			0.9 mg/kg (whole body) ⁴	Mortality, LOED				[20]	L; survival of fry reduced
Leuciscus cephalus cabeda, Chub	478 μg/kg OC		378 μg/kg lipid				0.790	[21,27]	F; %lipid = 1.27; %sed OC = 2.76
Alburnus alburnus alborella, Bleak	478 μg/kg OC		769 μg/kg lipid				1.608	[21,27]	F; %lipid = 1.95; %sed OC = 2.76
Cyprinus carpio, Carp	0.023 mg/kg n = 1	water = 0.85 ng/L n = 1	$0.48 \pm 0.26 \text{ mg/kg}$ n = 9					[14]	F; lower Detroit River; value is mean ± SD
Pimephales promelas, Fathead minnow			0.6 mg/kg (whole body) ⁴	Reproduction, LOED				[17]	L; significantly different from control (p = 0.05)
Gambusia affinis, Mosquito fish			5.3 mg/kg (whole body) ⁴	Mortality, NOED				[19]	L; no effect on survivorship after 3 days
Catastoma macrocheilus, Largescale sucker	530 μg/kg OC		1,261 µg/kg lipid				2.379	[23]	F; %lipid = 11.1; %sed OC = 1.0

Species:	Concentration	ncentration, Units in¹:			Ability to Accumulate²:			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Cottus cognatus, Slimy sculpin	2667 μg/kg OC		587.5 μg/kg lipid				0.220	[24]	F; %lipid = 8; %sed OC = 2.7	
		0.000093 μg/L	47 μg g/kg			5.70		[24]	F; %lipid = 8	
Comephorus dybowskii, Pelagic sculpin		particulate: 1.0 pg/L ±1.0 n = 7	0.12-0.16 mg/kg lipid (whole body) n = 1					[15]	F; Lake Baikal, Siberia	
		dissolved: 17 pg/L ±7.3 n = 7								
Wildlife										
Bucephala clangula, Goldeneye	0.023 mg/kg n = 1	water = 0.85 ng/L n = 1	0.080 ± 0.024 mg/kg $n = 3$					[14]	F; lower Detroit River; value is mean ± SD	
Aythya affinis, Lesser scaup	0.023 mg/kg n = 1	water = 0.85 ng/L n = 1	$0.093 \pm 0.027 \text{ mg/kg}$ n = 7					[14]	F; lower Detroit River; value is mean ± SD	
Aythya marila, Greater scaup	0.023 mg/kg n = 1	water = 0.85 ng/L n = 1	$0.14\pm0.045 \text{ mg/kg}$ n = 3					[14]	F; lower Detroit River; value is mean ± SD	

Species:	Concentration, Units in ¹ :			Toxicity:	Ability to Accumulate ² :			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Falco peregrinus, Peregrine falcon (eggs)			9 ng/g (eggs) n = 6	11.4% eggshell thinning				[16]	F; Kola Penninsula, Russia; n = number of clutches sampled	
Phoca siberica, Baikal seal		particulate: 1.0 pg/L ± 1.0 n = 7	2.0-2.2 mg/kg 5 lipid (blubber) n = 1					[15]	F; Lake Baikal, Siberia	
		dissolved: 17 pg/L ± 7.3 n = 7								

¹ Concentration units based on wet weight, unless otherwise noted.

² BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ This entry was excerpted directly from the Environmental Residue-Effects Database (ERED, www.wes.army.mil/el/ered, U.S. Army Corps of Engineers and U.S. Environmental Protection Agency). The original publication was not reviewed, and the reader is strongly urged to consult the publication to confirm the information presented here.

⁵ Not clear from reference if concentration is based on wet or dry weight.

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CASRN: 72-55-9

Chemical Category: PESTICIDE (ORGANOCHLORINE)

Chemical Name (Common Synonyms):

1,1'-(DICHLOROETHYLIDENE)BIS(4-CHLOROBENZENE),

p,p'-DICHLORODIPHENYLDICHLOROETHYLENE

4,4'-DICHLORODIPHENYLDICHLOROETHYLENE

Chemical Characteristics

Solubility in Water: 0.065 mg/L at 24 °C [1] **Half-Life:** 2.0 - 15.6 years based on

biodegradation of DDD in aerobic soils under field conditions [2]

Log K_{ow}: 6.76 [3] **Log K_{oc}:** 6.65 L/kg organic carbon

Human Health

Oral RfD: No data [4] Confidence: —

Critical Effect: Liver tumors in mice and hamsters, thyroid tumors in female rats

Oral Slope Factor: 3.4 x 10⁻¹ per (mg/kg)/day [4] Carcinogenic Classification: B2 [4]

Wildlife

Partitioning Factors: Based on the data presented in one study, log BCFs for birds collected from the lower Detroit River ranged from 5.92 to 6.36. Concentrations of DDE in birds were 40 to 108 times higher than in sediment. BSAFs were calculated for red-winged blackbird eggs and tree swallow eggs during a study in the Great Lakes area, with values ranging from 13 to 870 as reported in the attached summary table. BSAFs for tree swallow nestlings were 5 and 49.

Food Chain Multipliers: Biomagnification factors of 3.2 and 85 were determined for DDT and DDE, respectively, from alewife to herring gulls in Lake Ontario [5]. A study of arctic marine food chains measured biomagnification factors for DDE that ranged from 17.6 to 62.2 for fish to seal, 0.3 to 0.7 for seal to bear, and 10.7 for fish to bear [6].

Aquatic Organisms

Partitioning Factors: Partitioning factors for DDE in aquatic organisms were not calculated in the studies reviewed. However, the data showed ratios of DDT in tissue to sediment of 0.49 for oligochaetes and 32 for fish from the lower Detroit River. Ratios of DDT in lipid to sediment for three fish species from Rio de la Plata, Argentina ranged from 87 to 26,000. BSAFs for clams ranged from 1.2313 to

107.7 [15,41,36]. BSAFs for dover sole collected in southern California ranged from 1.7 to 3.4. BSAFs for other species ranged from 1.274 to 140.

Food Chain Multipliers: Food chain multipliers (FCMs) for trophic level 3 aquatic organisms were 23.7 (all benthic food web), 1.7 (all pelagic food web), and 14.4 (benthic and pelagic food web). FCMs for trophic level 4 aquatic organisms were 57.5 (all benthic food web), 3.7 (all pelagic food web), and 26.7 (benthic and pelagic food web) [46].

Toxicity/Bioaccumulation Assessment Profile

DDT is very persistent in the environment due to its low vapor pressure, high fat solubility, and resistance to degradation and photooxidation. DDT is degraded to DDE under aerobic conditions and to DDD in anoxic systems [7]. These metabolites, DDD and DDE, are similar to DDT in both their stability and toxicity. Chronic effects of DDT and its metabolites on ecological receptors include changes in enzyme production, hormonal balance, and calcium metabolism, which may cause changes in behavior and reproduction. The high octanol-water partition coefficient of DDT indicates that it is easily accumulated in tissues of aquatic organisms. Laboratory studies have shown that these compounds are readily bioconcentrated in aquatic organisms, with reported log BCFs for DDT ranging from 3.08 to 6.65 and for DDE ranging from 4.80 to 5.26 [8].

Invertebrate species are generally more susceptible than fish species to effects associated with exposure to DDT in the water column [8]. In general, the low solubility of DDT and its metabolites in water suggests that water column exposures are likely to be lower than exposures from ingestion of food or sediment. Sediments contaminated with pesticides, including DDT, have been shown to impact benthic communities at low concentrations. Results of laboratory and field investigations suggest that chronic effects generally occur at total DDT concentrations in sediment exceeding 2 μ g/kg [9]. Equilibrium partitioning methods predict that chronic effects occur at DDT concentrations in sediment of 0.6 to 1.7 μ g/kg [10].

For fish, the primary route of uptake is via prey items, but both DDT and its metabolites can be accumulated through the skin or gills upon exposure to water. Short-term exposure to DDT concentrations of less than $1 \mu g/L$ have been reported to elicit toxic responses in both freshwater and marine fish [8]. DDT may also be transferred to embryos from contaminated adults. DDT concentrations of 1.1 to 2.4 mg/kg in fish embryos have been associated with fry mortality [11,12].

Eggshell thinning, embryo mortality, and decreased hatchling survival have been linked to chronic exposure to DDT and its metabolites in the diet of birds. Of the three compounds, evidence strongly indicates that DDE is responsible for most reproductive toxicity in avian species [13]. Measurements of residues in eggs of birds are a reliable indicator of adverse effects. There is a large amount of variability in sensitivity to DDT and its metabolites among bird species, with waterfowl and raptor species showing the greatest sensitivities. Studies have shown the brown pelican to be most susceptible to adverse effects, with eggshell thinning and depressed productivity occurring at $3.0 \,\mu\text{g/g}$ of DDE in the egg and total reproductive failure when residues exceed $3.7 \,\mu\text{g/g}$ [13].

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Invertebrates									
Tubifex sp., Oligochaetes	0.012 mg/kg (n = 1)	Surface water 0.57 ng/L (n = 1)	0.0059 mg/kg (n = 1)					[14]	F; lower Detroit River; value is mean ± SD
Viviparus conectus, Gastropod mollusk	294 μg/kg OC		368 μg/kg lipid				1.2517	[36]	F; %lipid = 7.06; %sed OC = 1.02
Unio elongatulus, Bivalve mollusk	294 μg/kg OC		362 μg/kg lipid				1.2313	[36]	F; %lipid = 10.49; %sed OC = 1.02
Mollusks (unspecified)	99.67 μg/kg OC		229 μg/kg lipid				2.298	[37]	F; %lipid = 1.1; %sed OC = 2.8
Macomona liliana, Mollusk	36.67 μg/kg OC		522.20 μg/kg lipid				14.241	[38]	F; %lipid = 2.95; %sed OC = 0.30
	35.62 μg/kg OC		573.39 μg/kg lipid				16.097	[38]	F; %lipid = 2.33; %sed OC = 0.73
	36.36 μg/kg OC		278.21 µg/kg lipid				7.652	[38]	F; %lipid = 2.57; %sed OC = 0.22
	20 μg/kg OC		328.92 µg/kg lipid				16.446	[38]	F; %lipid = 2.04; %sed OC = 0.25

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	to Accumu	cumulate ² : Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
	6.25 μg/kg OC		61.34 μg/kg lipid				9.814	[38]	F; %lipid = 3.13; %sed OC = 0.48
Austrovenus stutchburyi, Mollusk	36.67 μg/kg OC		141.64 μg/kg lipid				3.863	[38]	F; %lipid = 5.62; %sed OC = 0.30
	35.62 μg/kg OC		148.75 μg/kg lipid				4.176	[38]	F; %lipid = 5.21; %sed OC = 0.73
	36.36 μg/kg OC		57.94 μg/kg lipid				1.594	[38]	F; %lipid = 4.85; %sed OC = 0.22
	20 μg/kg OC		59.95 μg/kg lipid				2.998	[38]	F; %lipid = 3.87; %sed OC = 0.25
	6.25 μg/kg OC		10.54 μg/kg lipid				1.686	[38]	F; %lipid = 4.27; %sed OC = 0.48
Corbicula fluminea, Asian clam	13 μg/kg OC		1,400 μg/kg lipid				107.7	[15]	F; %lipid not reported; %sed OC = 2.3
Corbicula fluminea, Asian clam	(0-5 cm) 0.3 ng/g dw	Surface water 1.8 ng/L	1.4 μg/g lipid (whole tissue)					[15]	F; Rio de La Plata, Argentina; lipid content 2.4-3.8%
	0.6 ng/g dw		1.4 μg/g lipid (whole tissue)						
Corbicula fluminea, Asian clam	9,664 μg/kg OC		540,984 µg/kg lipid				55.979	[41]	F; %lipid = 0.61; %sed OC = 0.19
	168 μg/kg OC		4,098 μg/kg lipid				24.393	[41]	F; %lipid = 0.61; %sed OC = 0.19

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability	to Accumul	mulate ² : Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
	210 μg/kg OC		2,131 µg/kg lipid				10.148	[41]	F; %lipid = 0.61; %sed OC = 0.19
Astacidae, Crayfish	99.67 μg/kg OC		177 μg/kg lipid				1.776	[37]	F; %lipid = 1.3; %sed OC = 2.8
Chironomus riparius, Midge			1.6 mg/kg (whole body) ⁴	Behavior, NOED				[31]	L; no effect on swimming behavior
			0.27 mg/kg (whole body) ⁴	Behavior, NOED				[31]	L; no effect on swimming behavior
			0.1 mg/kg (whole body) ⁴	Behavior, NOED				[31]	L; no effect on swimming behavior
Chironomus riparius, Midge			7.35 mg/kg (whole body) ⁴	Development, LOED				[34]	L; development time from egg to 4th instar decreased from 22- 25 days to 19-21 days
			3.75 mg/kg (whole body) ⁴	Development, NOED				[34]	L; no effect on developmental period of larvae
Fishes									
Anguilla anguilla, Eel	5 μg/kg OC		156 μg/kg lipid				31.200	[43]	F; %lipid = 7; %sed OC = 7

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability 1	to Accumul	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
	5 μg/kg OC		213 μg/kg lipid				42.600	[43]	F; %lipid = 7; %sed OC = 14
	32 μg/kg OC		2117 μg/kg lipid				66.156	[43]	F; %lipid = 6; %sed OC = 18
	76 μg/kg OC		849 µg/kg lipid				11.171	[43]	F; %lipid = 10; %sed OC = 12
	23 μg/kg OC		658 µg/kg lipid				28.609	[43]	F; %lipid = 10; %sed OC = 12
	72 μg/kg OC		2,176 µg/kg lipid				30.222	[43]	F; %lipid = 13; %sed OC = 32
Oncorhynchus mykiss, Rainbow trout			0.15 mg/kg (fat) ⁴	Growth, ED40				[29]	L; 40% decrease in growth relative to control
			0.15 mg/kg (fat) ⁴	Physiological, ED30				[29]	L; 30% decrease in hemoglobin content relative to control
			0.15 mg/kg (fat) ⁴	Physiological, ED30				[29]	L; 30% increase in liver size relative to control
			0.08 mg/kg (fat) ⁴	Physiological, ED35				[29]	L; 35% increase in kidney size relative to control

Species:	Concentration, Units in ¹ :		Toxicity:	Ability t	o Accumul	ate ² :	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Oncorhynchus, Salmo, Salvelinus spp., Salmonids	1,889 μg/kg OC		7,817 μg/kg lipid				4.139	[40]	F; %lipid = 11; %sed OC =2.7
		0.000076 μg/L	860 μg/kg			7.05		[40]	F; %lipid = 11
Oncorhynchus sp., Salmon	99.67 μg/kg OC		925 μg/kg lipid				9.281	[37]	F; %lipid = 13.1; %sed OC = 2.8
Prosopium williamsoni, Mountain whitefish	544.4 μg/kg OC		2,333 µg/kg lipid				4.285	[39]	F; %lipid = 12.0, %sed OC = 0.9
	3,500 μg/kg OC		4,460 µg/kg lipid (arithmetic mean of two samples)				1.274	[39]	F; %lipid = 12.25 %sed OC = 0.3
Coregonus autumnalis, Omul (endemic whitefish)		particulate: <14 pg/L n = 7	0.31-0.50 mg/kg lipid n = 2						
		dissolved: 17±7.1 pg/L n = 7							

Species:	Concentration, Units in 1:		Toxicity:	Ability to Accumulate ² :			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Salvelinus fontinalis, Brook trout			44.9 mg/kg (whole body) ⁴	Behavior, LOED				[30]	L; temperature selection after 24 h exposure to chemical
Salvelinus namaycush, Lake trout			1.09 mg/kg (whole body) ⁴	Mortality, LOED				[33]	L; survival of fry reduced
Alburnus alburnus alborella, Bleak fish	294 μg/kg OC		1,092 μg/kg lipid				3.7143	[36]	F; %lipid = 21.43; %sed OC = 1.02
Alburnus alburnus alborella, Bleak fish	358 μg/kg OC		2,113 μg/kg lipid				5.9022	[35, 45]	F; %lipid = 1.95; %sed OC = 2.76
Chondrostoma soetta	294 μg/kg OC		1,179 μg/kg lipid				4.0102	[36]	F; %lipid = 9.75; %sed OC = 1.02
Cyprinus carpio, Common carp	99.67 μg/kg OC		4,209 μg/kg lipid				42.229	[37]	F, %lipid = 13.9; %sed OC = 2.8
Cyprinus carpio, Common carp	174 μg/kg OC		1,905 µg/kg lipid				10.948	[42]	F, %lipid = 8.4; %sed OC = 2.13

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Cyprinus carpio, Carp	0.012 mg/kg (n = 1)	Surface water 0.57 ng/L (n = 1)	$0.38 \pm 0.15 \text{ mg/kg}$ (n = 9)					[14]	F; lower Detroit River; value is mean ± SD
Scardinius erythrophalmus, Rudd	294 μg/kg OC		1,473 μg/kg lipid				6.546	[36]	F; %lipid = 11.66; %sed OC = 1.02
Leuciscus cephalus, Chub	294 μg/kg OC		1,473 µg/kg lipid				5.0102	[36]	F; %lipid = 9.98; %sed OC = 1.02
Leuciscus cephalus cabeda, Chub	358 μg/kg OC		1,953 μg/kg lipid				5.4553	[35, 45]	F; %lipid = 1.27; %sed OC = 2.76
Rutilus pigus	294 μg/kg OC		728 μg/kg lipid				2.4762	[36]	F; %lipid = 12.63; %sed OC = 1.02
Rutilus rubilio	294 μg/kg OC		1,167 µg/kg lipid				3.9694	[36]	F; %lipid = 11.05; %sed OC = 1.02
Catostomus commersoni, White sucker	208 μg/kg OC		1,519 µg/kg lipid				7.303	[42]	F; %lipid = 7.9; %sed OC = 1.44

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
mixed <i>Catastoma</i> sp., Suckers	3,500 μg/kg OC		3,253 µg/kg lipid (arithmetic mean of two samples)				0.929	[39]	F; %lipid = 9.2; %sed OC = 0.3	
Catastoma macrocheilus, Largescale sucker	3,010 μg/kg OC		7,477 μg/kg lipid				2.484	[39]	F; %lipid = 11.1; %sed OC = 1.0	
Barbus barbus, Barbel	294 μg/kg OC		1,333 μg/kg lipid				4.5340	[36]	F; %lipid = 16.43; %sed OC = 1.02	
Siluris glanis, Wels fish, juveniles	294 μg/kg OC		731 µg/kg lipid				2.4864	[36]	F; %lipid = 3.83; %sed OC = 1.02	
Siluris glanis, Wels fish, adults	294 μg/kg OC		1,613 µg/kg lipid				5.4864	[36]	F; %lipid = 5.38; %sed OC = 1.02	
Pimelodus albicans, Mandi	0.2 ng/g dw		0.6 μg/g lipid (n = 2) (muscle)					[15]	F; Rio de La Plata, Argentina; lipid content 4%	
Pimelodus albicans, Mandi	20 μg/kg OC		600 μg/kg lipid				30.0	[15]	F; %lipid not reported; %sed OC = 1.0	
Gambusia affinis, Mosquito fish			29.2 mg/kg (whole body) ⁴	Mortality, NOED				[32]	L; no effect on survivorship after 3 days	

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability to Accumulate ² :			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Ambloplites rupestris, Rock bass	99.67 μg/kg OC		365 μg/kg lipid				3.662	[37]	F; %lipid = 0.7; %sed OC = 2.8	
Sunfish (unspecified)	99.67 μg/kg OC		254 μg/kg lipid				2.548	[37]	F; %lipid = 3.7; %sed OC = 2.8	
Roccus chrysops, White bass	99.67 μg/kg OC		1,586 µg/kg lipid				15.913	[37]	F; %lipid = 1.8; %sed OC = 2.8	
Micropterus salmoides, Smallmouth bass	99.67 μg/kg OC		1,352 µg/kg lipid				13.565	[37]	F; %lipid = 0.6; %sed OC = 2.8	
Dorosoma cepedianum, Gizzard shad	99.67 μg/kg OC		382 μg/kg lipid				3.833	[37]	F; %lipid = 6.8; %sed OC = 2.8	
Perca fluviatilis, Perch	294 μg/kg OC		3,390 µg/kg lipid				11.5306	[36]	F; %lipid = 5.84; %sed OC = 1.02	
Stizostedion vitreum, Walleye	99.67 μg/kg OC		2,593 µg/kg lipid				26.016	[37]	F; %lipid = 1.2; %sed OC = 2.8	

Species:	Concentration, Units in ¹ :			Toxicity:	Abilit	y to Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Microstomus pacificus, Dover sole	27 μg/g dw		$16.0 \mu g/g (n=5)$			log MBAF -0.26	1.7	[16]	F; Southern California Bight; modified
sole	(n = 5)		(muscle)						bioaccumulation
			210 μ g/g (n = 3) (liver)			0.79	2.0		factor (MBAF) = C_{org} ww/ C_{sed} dw; water content of tissue was not measured
						log MBAF			
	$0.09 \mu g/g$ dw (n = 10)		$0.24 \mu g/g \text{ (n = 10)}$ (muscle)			0.43	1.8		
	O. (L. 13)		$0.80 \mu g/g (n = 6)$ (liver)			1.79	3.4		
Oligosarcus jenynsi, Common name not available	5.7 ng/g dw		0.5 μ g/g lipid (n = 7) (muscle)					[15]	F; Rio de La Plata, Argentina; lipid content 0.32%
Prochilodus platensis, Curimata	20 μg/kg OC		2,800 µg/kg lipid				140	[15]	F, %lipid not reported; %sed OC = 1.0
Prochilodus platensis, Curimata	0.2 ng/g dw		Three composite samples: 1.2 (n = 4), 5.2 (n = 4) and 2 (n = 5) μ g/g lipid (muscle)					[15]	F; Rio de La Plata, Argentina; lipid content 1-12.7%

Species:	Concentration, Units in ¹ :			Toxicity:	Ability to Accumulate ² :			Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Gar pike (unspecified)	99.67 μg/kg OC		11,986 μg/kg lipid				120.257	[37]	F; %lipid = 0.8; %sed OC = 2.8
Comephorus bybowskii, Pelagic sculpin,		particulate: <14 pg/L n = 7 dissolved: 17 pg/L ± 7.1 n = 7	0.74-0.76 mg/kg lipid n = 1					[17]	F; Lake Baikal, Siberia
Cottus cognatus, Slimy sculpin	1,889 μg/kg OC		2,375 µg/kg lipid				1.257	[40]	F; %lipid = 8; %sed OC = 2.7
		0.000076 μg/L	190 μg/kg			6.40		[40]	F; %lipid =8; %sed OC = 2.7
Wildlife									
Bucephala clangula, Goldeneye	0.012 mg/kg (n = 1) seston = 0.10 mg/kg	Surface water 0.57 ng/L (n = 1)	0.48 ± 0.18 mg/kg (whole body) (n = 3)					[14]	F; lower Detroit River; value is mean ± SD

Species:	Concentration	Concentration, Units in ¹ :			Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Aythya affinis, Lesser scaup	0.012 mg/kg (n = 1)	Surface water 0.57 ng/L (n = 1)	0.80 ± 0.33 mg/kg (whole body) (n = 7)					[14]	F; lower Detroit River; value is mean ± SD
Aythya marila, Greater scaup	0.012 mg/kg (n = 1)	Surface water 0.57 ng/L (n = 1)	$1.3 \pm 0.25 \text{ mg/kg}$ (whole body) (n = 3)					[14]	F; lower Detroit River; value is mean ± SD
Falco peregrinus, Peregrine falcon			µg/g (egg): ≤15 15-30 >30	Young produced per active nest: 1.8 2.0 1.0				[26]	F; Alaska; young produced not adjusted for sample egg collected

Species:	Concentrat	tion, Units in ¹ :	Toxicity:	Ability	to Accumu	ılate²:	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Aquila chrysaetos, Golden eagle			μg/g (egg): 0.1 0.1 0.2 0.3	Mean % eggshell thinning = 7% 1% 3% 4% 5%				[24]	F; Great Britain; percentage of thinning based on thickness index [24]
Haliaeetus leucocephalus, Bald eagle			10 μg/g (egg)	Mean percent eggshell thinning= 10%				[22]	F; Oregon and Washington
			µg/g (egg): <2.2 2.2-3.5 3.6-6.2 6.3-11.9 ≥12	Young produced per active nest: 1.0 1.0 0.5 0.3 0.2				[23]	F
Ardea herodias, Great blue heron			4 μg/g (egg) 5 μg/g (egg)	Mean percent eggshell thinning = 10% 13%				[18]	F; Washington

Species:	Concentrat	tion, Units in¹:		Toxicity:	Ability	to Accumi	ulate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Plegadis chihi, White-faced ibis			2 μg/g (egg) 1 μg/g (egg)	Mean percent eggshell thinning= 12% 8%				[20]	F; Nevada
			µg/g (egg): ≤1 1-4 4-8 8-16 >16	Young produced per active nest: 1.8 1.8 1.3 0.8 0.6				[21]	F; Nevada; young produced not adjusted for sample egg collected
Egretta thula, Snowy egret			1 μg/g (egg) 2 μg/g (egg)	Mean percent eggshell thinning= 3% 12%				[20]	F; Nevada; young produced not adjusted for sample egg collected
			µg/g (egg): ≤1 1-5 5-10 10-20	Young produced per active nest: 2.2 2.4 1.0 1.0					

Species:	Concentrat	ion, Units in¹:	Toxicity:	Ability 1	to Accumu	late²:	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Sula bassanus, Northern gannet			μg/g (egg) 19	Mean % eggshell thinning = 17%				[19]	F; Quebec
Larus californicus, California gull			430 mg/kg (brain) ⁴	Mortality, not available (NA)				[28]	L
			175 mg/kg (breast) ⁴	Mortality, NA				[28]	L
			3,100 mg/kg (liver) ⁴	Mortality, NA				[28]	L
			220 mg/kg (brain) ⁴	NA, NA				[28]	L
			490 mg/kg (breast) ⁴	NA, NA				[28]	L
			800 mg/kg (liver) ⁴	NA				[28]	L
			750 mg/kg (liver) ⁴	NA				[28]	L
Pelecanus			4.4 mg/kg (brain) ⁴	Mortality, NA				[28]	L
occidentalis, Brown pelican			59.5 mg/kg (breast) ⁴	Mortality, NA				[28]	L
			7.15 mg/kg (liver) ⁴	Mortality, NA				[28]	L

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Agelaius phoeniceus, Red-	40.5 ng/g TOC = 2.5%		3,088.1 ng/g				41.4	[25]	F; Great Lakes/St. Lawrence River
winged blackbird (eggs)	7.9 ng/g TOC = 21.0%		777.7 ng/g				372.7		basin; 12 wetlands sites; sediment concentration reported as wet
	373.1 ng/g TOC = 7.5%		648.7 ng/g				12.9		weight concentration which may be a
	1,160.7 ng/g TOC = 12%		1,299.6 ng/g				13.2		typographical error
	10.4 ng/g TOC-18.5%		305.7 ng/g				113.3		
	65.4 ng/g TOC = 11.5%		826.2 ng/g				30.3		
	1.6 ng/g TOC = 10.5%		416.1 ng/g	582.4					
	0.8 ng/g TOC = 13.8%		145.1 ng/g				522		
	1.3 ng/g TOC =		183.5 ng/g				326.4		
	11.1%						203.7		
	3.0 ng/g TOC = 23.9%		117.6 ng/g						

Species:	Concentrat	tion, Units in¹:		Toxicity:	Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Tachycineta bicolor, Tree swallow Nestlings	65.4 ng/g TOC = 11.5% 0.8 ng/g TOC =		(whole body minus feet, beak, wings, and feathers) 288.2 ng/g				548.9	[25]	F; Great Lakes/St. Lawrence River basin; 12 wetlands sites; sediment concentration reported as wet weight concentration which may be a
Eggs	13.8% 65.4 ng/g TOC = 11.5%		794.7 ng/g				16.2		typographical error
	0.8 ng/g TOC = 13.8%		458.2 ng/g				868.6		
			$3.5 \mu g/g \text{ (egg)}$ $(n = 6)$	11.4% eggshel	11			[27]	F; Kola Penninsula, Russia; n = number of clutches sampled

Species:	Concentrat	Concentration, Units in ¹ :			Ability to Accumulate ² :			Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Phoca siberica, Baikal seal		particulate: <14 pg/L n = 7	43-44 mg/kg lipid n = 1					[17]	F; Lake Baikal, Siberia
		dissolved: 17 pg/L ± 7.1 n = 7							

Concentration units based on wet weight unless otherwise noted.
 BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ This entry was excerpted directly from The Environmental Residue-Effects Database (ERED, www.wes.army.mil/el/ered, U.S. Army Corps of Engineers and U.S. Environmental Protection Agency). The original publication was not reviewed, and the reader is strongly urged to consult the publication to confirm the information presented here.

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Chemical Category: PESTICIDE (ORGANOCHLORINE)

Chemical Name (Common Synonyms):

CASRN: 50-29-3

1,1'-(2,2,2-TRICHLOROETHYLIDENE)BIS(4-CHLOROBENZENE),

p,p′-DICHLORODIPHENYLTRICHLOROETHANE, 4,4′-DICHLORODIPHENYLTRICHLOROETHANE

Chemical Characteristics

Solubility in Water: 0.0031 - 0.0034 mg/L **Half-Life:** 2.0 - 15.6 years based on

at 25°C [1] biodegradation of DDD in

aerobic soils under field

conditions [2]

Log K_{ow}: 6.83 [3] **Log K_{oc}:** 6.71 L/kg organic carbon

Human Health

Oral RfD: 5 x 10⁻⁴ mg/kg/day [4] Confidence: Medium, uncertainty factor

= 100

Critical Effect: Liver lesions in rats, liver tumors in mice and rats

Oral Slope Factor: 3.4 x 10⁻¹ per (mg/kg)/day [4] Carcinogenic Classification: B2 [4]

Wildlife

Partitioning Factors: Partitioning factors for DDT in wildlife were not calculated in the studies reviewed. However, based on the data in one study, log BCFs for birds from the lower Detroit River ranged from 4.81 to 5.01. Concentrations of DDT in birds were 2.1 to 3.3 times higher than in sediment.

Food Chain Multipliers: Biomagnification factors of 3.2 and 85 were determined for DDT and DDE, respectively, from alewife to herring gulls in Lake Ontario [5]. A study of arctic marine food chains measured biomagnification factors for DDE that ranged from 17.6 to 62.2 for fish to seal, 0.3 to 0.7 for seal to bear, and 10.7 for fish to bear [6].

Aquatic Organisms

Partitioning Factors: Based on the results from one study reviewed, the log BCF for carp collected from the lower Detroit River was 4.77. Ratios of DDT in lipids to sediment were 450 in clams and 1,250 to 11,000 in fish from the Rio de la Plata, Argentina. BSAFs for clams ranged from 0.060 to 302.326 [14,33,36]. BSAFs for fish ranged from 0.120 to 88.07.

Food Chain Multipliers: Food chain multipliers (FCMs) for trophic level 3 aquatic organisms were 22.5 (all benthic food web), 1.7 (all pelagic food web), and 13.7 (benthic and pelagic food web). FCMs for trophic level 4 aquatic organisms were 52.5 (all benthic food web), 3.6 (all pelagic food web), and 24.6 (benthic and pelagic food web) [39].

Toxicity/Bioaccumulation Assessment Profile

DDT is very persistent in the environment due to its low vapor pressure, high fat solubility, and resistance to degradation and photooxidation. DDT is degraded to DDE under aerobic conditions and to DDD in anoxic systems [7]. These metabolites, DDD and DDE, are similar to DDT in both their stability and toxicity. Chronic effects of DDT and its metabolites on ecological receptors include changes in enzyme production, hormonal balance, and calcium metabolism, which may cause changes in behavior and reproduction. The high octanol-water partition coefficient of DDT indicates that it is easily accumulated in tissues of aquatic organisms. Laboratory studies have shown that these compounds are readily bioconcentrated in aquatic organisms, with reported log BCFs for DDT ranging from 3.08 to 6.65 and for DDE ranging from 4.80to 5.26 [8].

Invertebrate species are generally more susceptible than fish species to effects associated with exposure to DDT in the water column [8]. In general, the low solubility of DDT and its metabolites in water suggests that water column exposures are likely to be lower than exposures from ingestion of food or sediment. Sediments contaminated with pesticides, including DDT, have been shown to affect benthic communities at low concentrations. Results of laboratory and field investigations suggest that chronic effects generally occur at total DDT concentrations in sediment exceeding 2 μ g/kg [9]. Equilibrium partitioning methods predict that chronic effects occur at DDT concentrations in sediment of 0.6 to 1.7 μ g/kg [10].

For fish, the primary route of uptake is via prey items, but both DDT and its metabolites can be accumulated through the skin or gills upon exposure to water. Short-term exposure to DDT concentrations of less than $1 \mu g/L$ have been reported to elicit toxic responses in both freshwater and marine fish [8]. DDT may also be transferred to embryos from contaminated adults. DDT concentrations of 1.1 to 2.4 mg/kg in fish embryos have been associated with fry mortality [11,12].

Eggshell thinning, embryo mortality, and decreased hatchling survival have been linked to chronic exposure to DDT and its metabolites in the diet of birds. Of the three compounds, evidence strongly indicates that DDE is responsible for most reproductive toxicity in avian species [13]. Measurements of residues in eggs of birds are a reliable indicator of adverse effects. There is a large amount of variability in sensitivity to DDT and its metabolites among bird species, with waterfowl and raptor species showing the greatest sensitivities. Studies have shown the brown pelican to be most susceptible to adverse effects, with eggshell thinning and depressed productivity occurring at 3.0 μ g/g of DDE in the egg and total reproductive failure when residues exceed 3.7 μ g/g [13].

Species:	Concentration	on, Units in¹:	Toxicity:	Ability	to Accumu	late ² :	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Invertebrates									
Macomona liliana, Mollusk	63.33 μg/kg OC		13.22 μg/kg lipid				0.209	[33]	F; %lipid = 2.95; %sed OC = 0.30
	76.71 μg/kg OC		33.91 µg/kg lipid				0.442	[33]	F; %lipid = 2.33; %sed OC = 0.73
	127.27 μg/kg OC		24.12 μg/kg lipid				0.190	[33]	F; %lipid = 2.57; %sed OC = 0.22
	20.83 μg/kg OC		7.35 μg/kg lipid				0.353	[33]	F; %lipid = 3.13; %sed OC = 0.48
Austrovenus stutchburyi, Mollusk	63.33 μg/kg OC		8.01 μg/kg lipid				0.126	[33]	F; %lipid = 5.62; %sed OC = 0.30
	76.71 μg/kg OC		7.29 µg/kg lipid				0.095	[33]	F; %lipid = 5.21; %sed OC = 0.73
	127.71 μg/kg OC		7.63 μg/kg lipid				0.060	[33]	F; %lipid = 4.85; %sed OC = 0.22
Corbicula fluminea, Asian clam	4.3 μg/kg OC		1,300 µg/kg lipid				302.326	[14]	F; %lipid = not reported; %sed OC = 2.3
Corbicula fluminea, Asian clam	3277 μg/kg OC		108,197 μg/kg lipid				33.017	[36]	F; %lipid = 0.61; %sed OC = 1.19

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability to	o Accumula	ite²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
	67 μg/kg OC		508 μg/kg lipid				7.563	[36]	F; %lipid = 0.61; %sed OC = 1.19
	92 μg/kg OC		164 μg/kg lipid				1.774	[36]	F; %lipid = 0.61; %sed OC = 1.19
Corbicula fluminea, Asian clam	(0-5 cm) 2.9 ng/g dw		1.3 μg/g lipid (whole tissue)					[14]	F; Rio de La Plata, Argentina; lipid content 2.4-3.8%
Mercenaria mercenaria, Quahog clam			0.126 mg/kg (whole body) ⁴	Behavior, NOED				[29]	L; No effect on feeding activity
Mya arenaria, Soft shell clam				NOED				[29]	L; no effect on feeding activity
Daphnia magna, Cladoceran			1.83 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship in 3 days
Diporeia spp., Amphipod			19.7 mg/kg (whole body) ⁴	Mortality, NOED				[22]	L; no increase in mortality in 96 hours

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability t	o Accumula	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Gammarus fasciatus, Amphipod			0.336 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship in 3 days
Palaemonetes kadiakensis, Grass shrimp			0.1 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship in 3 days
Orconectes nais, Crayfish			0.0466 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship in 3 days
Ephemera danica, Mayfly			6 mg/kg (whole body) ⁴	Growth, NOED				[20]	L
			6 mg/kg (whole body) ⁴	Mortality, NOED					L
Hexagenia bilineata, Mayfly			0.336 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship after 3 days
Siphlonurus sp., Mayfly			0.216 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship after 3 days

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Libellula sp., Dragonfly			0.0144 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship after 2 days
Ischnura verticalis, Damselfly			0.075 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship after 2 days
Chironomus sp., Midge			0.44 mg/kg (whole body) ⁴	Mortality, NOED				[23]	L; no effect on survivorship in 3 days
Chironomus riparius, Midge			0.83 mg/kg (whole body) ⁴	Behavior, LOED				[24]	L; reduced swimming ability
			0.18 mg/kg (whole body) ⁴	Behavior, NOED				[24]	L; no effect on swimming behavior
			0.08 mg/kg (whole body) ⁴	Behavior, NOED				[24]	L; no effect on swimming behavior
Fishes									
Squalus acanthias, Spiny dogfish			0.1 mg/kg (whole body) ⁴	Mortality, NOED				[32]	L; no effect on mortality in 24 hours

Species:	Concentrat	centration, Units in¹:			Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Anguilla anguilla, Eel	23 μg/kg OC		158 μg/kg lipid				6.87	[37]	F; %lipid = 7; %sed OC = 7
	8 μg/kg OC		135 µg/kg lipid				16.88	[37]	F; %lipid = 7; %sed OC = 14
	14 μg/kg OC		1233 μg/kg lipid				88.07	[37]	F; %lipid = 6; %sed OC = 18
	25 μg/kg OC		221 µg/kg lipid				8.84	[37]	F; %lipid = 10; %sed OC = 12
	34 μg/kg OC		287 μg/kg lipid				8.44	[37]	F; %lipid = 10; %sed OC = 12
	144 μg/kg OC		1064 μg/kg lipid				7.39	[37]	F; %lipid = 13; %sed OC = 32
Oncorhynchus, Salmo, Salvelinus sp., Salmonids	667 μg/kg OC		727 μg/kg lipid				1.091	[35]	F; %lipid = 11; %sed OC = 2.7
		0.000019 μg/L	80 μg/kg lipid			6.62		[35]	F; %lipid = 11
Salmonids							1.67	[38]	F
Oncorhynchus kisutch, Coho salmon			95 mg/kg (whole body) ⁴	Mortality, ED50				[27]	L; 50% mortality in 31 days

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability t	o Accumul	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Prosopium williamsoni, Mountain whitefish	244 μg/kg OC		417 μg/kg lipid				1.706	[34]	F; %lipid = 12.0; %sed OC = 0.9
	6,433 μg/kg OC		772 μg/kg lipid				0.120	[34]	F; %lipid = 12.25; %sed OC = 0.30
Corogonus autumnalis, Omul (endemic whitefish)		particulate: 5.1pg/L ± 2.3 n = 7	$0.16-0.27 \text{ mg/kg}^5 \text{ lipid}$ (whole body) n = 2					[16]	F; Lake Baikal, Siberia
		dissolved: $50 \text{ pg/L} \pm 23$ n = 7							
Salmo salar, Atlantic salmon			3 mg/kg (whole body) ⁴	Morphology, NOED				[26]	L; no effect on metabolic rate or growth, resd_conc_wet value range 3.0-5.0
Salvelinus namaycush, Lake trout			3.9 mg/kg (whole body) ⁴	Behavior, LOED				[21]	L; temperature selection after 24- hour exposure to chemical

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	to Accumu	ılate²:	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
			27.8 mg/kg (whole body) ⁴	Behavior, LOED				[21]	L; temperature selection after 24- hour exposure to chemical	
Salvelinus namaycush, Lake trout			3.66 mg/kg (whole body) ⁴	Mortality, LOED				[28]	L; survival of fry reduced	
			2 mg/kg (whole body) ⁴	Mortality, LOED				[28]	L; survival of fry reduced	
Carassius auratus, Goldfish			5.1 mg/kg (whole body) ⁴	Behavior, LOED				[31]	L; behavioral changes, loss of equilibrium, convulsions	
Pimephales promelas, Fathead minnow			3.8 mg/kg (whole body) ⁴	Reproduction, LOED				[19]	L; significantly different from control $(p = 0.05)$	
			24 mg/kg (whole body) ⁴	Reproduction, LOED				[19]	L; significantly different from control (p = 0.05)	
Cyprinus carpio, Carp	0.012 mg/kg (n = 1)	Surface water 0.39 ng/L (n = 1)	$0.023 \pm 0.012 \text{ mg/kg}$ (n = 9)					[15]	F; lower Detroit River; value is mean ± SD	

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability to	o Accumula	ate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Mixed <i>Catastoma</i> sp., Suckers	6433 μg/kg OC		869 μg/kg lipid				0.135	[34]	F; %lipid = 9.2; %sed OC = 0.30
Catastoma macrocheilus, Largescale sucker	340 μg/kg OC		811 μg/kg lipid				2.385	[34]	F; %lipid = 11.1; %sed OC = 1.0
Pimelodus albicans, (Marine catfish)	0.4 ng/g dw		0.5 μ g/g lipid (n = 7) (muscle)					[14]	F; Rio de La Plata, Argentina; lipid content 4%
Pimelodus albicans, Mandi	40.0 μg/kg OC		500 μg/kg lipid					[14]	F; %lipid = not reported; %sed OC = 1.0
Gambusia affinis, Mosquito fish			18.6 mg/kg (whole body) ⁴	Mortality, NOED				[25]	L; no effect on survivorship after 3 days
Leuciscus idus, Golden ide			95 mg/kg (whole body) ⁴	Mortality, NOED				[30]	L; no effect on survivorship in 3 days

Species:	Concentrati	on, Units in¹:		Toxicity:	Ability 1	to Accumu	late ² :	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Lepomis macrochirus, Bluegill			4.2 mg/kg (whole body) ⁴	Behavior, LOED				[31]	L; behavioral changes, loss of equilibrium, convulsions	
Comephorus dybowskii, Pelagic sculpin		particulate: $5.1 \text{pg/L} \pm 2.3$ n = 7 dissolved: $50 \text{ pg/L} \pm 23$ n = 7	0.52-0.64 mg/kg lipid (whole body) n = 1					[16]	F; Lake Baikal, Siberia	
Cottus cognatus, Slimy sculpin	667 µg/kg OC		362 μg/kg lipid				0.544	[35]	F; %lipid = 8; %sed OC = 2.7	
		0.000019 μg/L	29 μg/kg lipid			6.18		[35]	F; %lipid = 8	
Prochilodus platensis, common name not available	0.4 ng/g dw		Three composite samples (μ g/g lipid): 2.4 (n = 4) (muscle) 9.3 (n = 4) (muscle) 4.3 (n = 5) (muscle)					[14]	F; Rio de La Plata, Argentina; lipid content 1-12.7%	
Prochilodus platensis, Curimata	40.0 μg/kg OC		5,333.33 µg/kg lipid					[14]	F; %lipid = not reported; %sed OC = 1.0	

Species:	Concentration	on, Units in¹:		Toxicity:	Ability t	o Accumula	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Wildlife									
Bucephala clangula, Goldeneye	0.012 mg/kg (n = 1)	water = 0.39 ng/L (n = 1)	0.040 mg/kg (whole body) (n = 3)					[15]	F; lower Detroit River; value is mean ± SD
Aythya affinis, Lesser scaup	0.012 mg/kg (n = 1)	surface water 0.39 ng/L (n = 1)	0.025 mg/kg (whole body) (n = 7)					[15]	F; lower Detroit River; value is mean ± SD
Aythya marila, Greater scaup	0.012 mg/kg (n = 1)	surface water 0.39 ng/L (n = 1)	0.040 ± 0.0094 mg/kg (whole body) (n = 3)					[15]	F; lower Detroit River; value is mean ± SD
Falco peregrinus, Peregrine falcon (eggs)			22 ng/g (eggs) (n = 6)	11.4% eggshell thinning				[17]	F; Kola Penninsula, Russia; n = number of clutches sampled
Larus californicus, California gull			440 mg/kg (brain) ⁴	Mortality, NA				[18]	L
			183 mg/kg (breast) ⁴	Mortality, NA				[18]	L
			3200 mg/kg (liver) ⁴	Mortality, NA				[18]	L

Species:	Concentrat	ion, Units in¹:		Toxicity:	Ability to	o Accumula	ate ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Pelecanus occidentalis,			4.55 mg/kg (brain) ⁴	Mortality, NA				[18]	L
Brown pelican			66 mg/kg (breast) ⁴	Mortality, NA				[18]	L
			7.8 mg/kg (liver) ⁴	Mortality, NA				[18]	L
Phalacrocorax			230 mg/kg (brain) ⁴	Mortality, NA				[18]	L
penicillatus, Brandts cormorant			500 mg/kg (breast) ⁴	Mortality, NA				[18]	L
			840 mg/kg (liver) ⁴	Mortality, NA				[18]	L
			810 mg/kg (Liver) ⁴	Mortality, NA				[18]	L
Phoca siberica, Baikal seal		particulate: 5.1pg/L ± 2.3 n = 1	17-21 mg/kg ⁵ lipid (blubber) n = 1					[16]	F; Lake Baikal, Siberia
		dissolved: $50 \text{ pg/L} \pm 23$ n = 1							

¹ Concentration units based on wet weight unless otherwise noted.

² BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ This entry was excerpted directly from the Environmental Residue-Effects Database (ERED, www.wes.army.mil/el/ered, U.S. Army Corps of Engineers and U.S. Environmental Protection Agency). The original publication was not reviewed, and the reader is strongly urged to consult the publication to confirm the information presented here.

⁵ Not clear from reference if concentration is based on wet or dry weight.

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Chemical Category: PESTICIDE (ORGANOPHOSPHATE)

Chemical Name (Common Synonyms): DIAZINON CASRN: 333-41-5

Chemical Characteristics

Solubility in Water: 0.004% at 20°C [1] **Half-Life:** No data [2]

Log K_{ov}: 3.70 [3] **Log K**_{oc}: 3.64 L/kg organic carbon

Human Health

Oral RfD: 9 x 10⁻⁴ mg/kg/day [4] Confidence: —

Critical Effect: Decreased cholinesterase activity

Oral Slope Factor: No data [4,5] Carcinogenic Classification: No data [4,5]

Wildlife

Partitioning Factors: Partitioning factors for diazinon in wildlife were not found in the literature.

Food Chain Multipliers: Food chain multipliers for diazinon in wildlife were not found in the literature.

Aquatic Organisms

Partitioning Factors: Partitioning factors for diazinon in aquatic organisms were not found in the literature. Log BCFs ranged from 0.69 to 1.23 (invertebrates) and from 1.59 to 2.90 (fishes).

Food Chain Multipliers: Food chain multipliers for diazinon in aquatic organisms were not found in the literature.

Toxicity/Bioaccumulation Assessment Profile

Diazinon is relatively toxic to aquatic organisms. The acute toxicity for aquatic invertebrates ranged from 0.9 μ g/L (48-h LC50) for *Daphnia pulex* [6] to 200 μ g/L (96-h LC50) for *Gammarus lacustris* [7], while chronic toxicity ranged from 0.27 μ g/L (30-d LC50) for *Gammarus pseudolimneaus* to 4.6 μ g/L (30-d LC50) for *Acroneuria lycorias* [8]. The maximum acceptable concentration (MATC) for diazinon based on the exposure with sheepshead minnows, was 0.47 μ g/L [5], and 3.2 μ g/L based on the exposure with fathead minnows [9].

The mode of toxic action of organophosphorus compounds is related to the inhibition of acetylcholinesterase in tissue of animals [10]. A representative of organophosphorus insecticides, diazinon shows species-selective toxicity in fish [11]. For example, diazinon was about 10 times more toxic to the guppy than to the zebra fish [12] and 22 times more potent to loach than killifish [10]. Both the guppy and zebra fish metabolized diazinon to 2-isopropyl-6-methyl-4-pyrimidinol (pyrimidinol). The species-specific oxidative transformation of diazinon or inhibition of acetylcholinesterase are responsible for the differences in diazinon toxicity. During the exposure of pretreated fish (guppies and zebra fish) to diazinon [13], the tissue concentration of pyrimidinol initially increased, then declined to very low levels. Keizer et al. [13] hypothesized that the toxicity of diazinon to guppy is due to its metabolism to a highly toxic metabolite, e.g., diazoxon whereas toxicity to zebra fish is related to bioaccumulation of the parent compound. Fish reached an apparent steady state after 48 hours [12] or 96 hours [14].

Diazinon was most rapidly excreted from the gallbladder followed by liver, muscle, and kidney [11]. The slow excretion rate from kidney was probably because diazinon was transported from all parts of the fish to the kidney before excretion [15]. The log BCFs for eels exposed to $56 \mu g/L$ of diazinon were 2.90 in liver, 3.20 in muscle, and 3.36 in gill tissue [16]. Diazinon elimination from the selected tissues was rapid; it was not detected in any tissue after 24-hour exposure in clean water [16]. The results of the study by Kanazawa [17] showed that the concentration of diazinon in tissue of the freshwater fish reached a maximum after 4 days and then decreased gradually. The uptake of diazinon by killifish was not influenced if the fish were exposed to the individual pesticide, or to a pesticide mixture [18].

Diazinon was identified as a major toxicant in municipal effluents [19], indicating persistence of this pesticide in the environment. According to Lee et al. [20], the toxicity of diazinon can be induced by dissolved organic materials.

Species:	Concentrati	ion, Units in¹:		Toxicity:	Ability	to Accumula	ite ² :	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
Invertebrates									
Cipangopoludina malleata, Pond snail		10 μg/L			0.77			[21]	L
Procambarus clarkii, Crayfish		10 μg/L			0.69			[21]	L
Indoplanorbis exustus, Red snail		10 μg/L			1.23			[21]	L
Fishes									
Pseudorasbora parva, Topmouth gudgeon		50 μg/L	11.3 ng/g		2.32			[22]	L
Anguilla anguilla, Eel		10 μg/L	80 ng/g (liver)		2.90			[16]	L
Anguilla anguilla, Eel		10 μg/L	160 ng/g (muscle)		2.90			[16]	L
Gnathopogon caerulescens, Willow shiner					2.39			[23]	F

Species:	Concentration	on, Units in¹:		Toxicity:	Ability	to Accumu	late ² :	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Pseudorasbora parva, Topmouth gudgeon					2.18			[23]	F	
Pseudorasbora parva, Motsugo		0.7 mg/kg	211 mg/kg (4-day) 17 mg/kg (30 day)	Bleeding, abnormal swimming	1.81			[17]	L	
Brachydanio rerio, Zebra fish	Ingestion		1,550 mg/kg (whole body) ⁴	Mortality, ED100				[13]	L; Lethal body burden	
Zacco slatypus, Pale chub					2.18			[23]	F	
Plecoglossus altivelis, Ayu sweetfish					1.79			[23]	F	
Cyprinodon variegatus, Sheepshead minnow		1.8 μg/L	0.26 mg/kg in 4d, 0.11 mg/kg in 7d, 0.31 mg/kg in 14d		2.17			[14]	L	
•		3.5 µg/L	0.38 mg/kg in 4d, 0.21 mg/kg in 7d, 0.49 mg/kg in 14d		2.17			[14]	L	
		6.5 μg/L	1.3 mg/kg in 4d, 0.5 mg/kg in 7d, 1.4 mg/kg in 14d		2.33			[14]	L	
Cyprinodon variegatus, Sheepshead minnow			0.3 mg/kg (whole body) ⁴	Morphology, LOED				[14]	L; body darkened, lateral curvature of body	

Species:	Concentrati	ion, Units in¹:		Toxicity:	Ability	to Accumu	ılate²:	Source:	
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			1.4 mg/kg (whole body) ⁴	Morphology, not applicable				[14]	L; body darkened, lateral curvature of body
			0.5 mg/kg (whole body) ⁴	Morphology, not applicable				[14]	L; body darkened, lateral curvature of body
			0.05 mg/kg (whole body) ⁴	Mortality, NOED				[14]	L; no effect on morphology or appearance
			1.4 mg/kg (whole body) ⁴	Mortality, NOED				[14]	L; no effect on mortality
			0.5 mg/kg (whole body) ⁴	Mortality, NOED				[14]	L; no effect on mortality
			0.3 mg/kg (whole body) ⁴	Mortality, NOED				[14]	L; no effect on mortality
			0.05 mg/kg (whole body) ⁴	Mortality, NOED				[14]	L; no effect on mortality
			0.05 mg/kg (whole body) ⁴	Physiological, LOED				[14]	L; inhibition of acetylcholinesterase activity
			1.4 mg/kg (whole body) ⁴	Physiological, NA				[14]	L; 71% inhibition of acetylcholinesterase activity
			0.5 mg/kg (whole body) ⁴	Physiological, NA				[14]	L; inhibition of acetylcholinesterase activity
			0.3 mg/kg (whole body) ⁴	Physiological, NA				[14]	L; inhibition of acetylcholinesterase activity
			1.4 mg/kg (whole body) ⁴	Reproduction, ED50				[14]	L; 45-55% reduction in average number of eggs produced

Species:	Concentrati	ion, Units in¹:		Ability to Accumulate ² :			Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³
			0.5 mg/kg (whole body) ⁴	Reproduction, ED50				[14]	L; 45-55% reduction in average number of eggs produced
			0.3 mg/kg (whole body) ⁴	Reproduction, ED50				[14]	L; 45-55% reduction in average number of eggs produced
			0.05 mg/kg (whole body) ⁴	Reproduction, LOED				[14]	L; statistically significant reduction in number of eggs produced
Poecilia reticulata, Guppy	,	0.8 mg/L	25.8 μg/g in 24h, 90.3 μg/g in 48h, 167.7 μg/g in 96h, 109 mg/kg (whole body) ⁴	Mortality, ED100				[13]	L; lethal body burden
Poecilia reticulata, Guppy			2,430 mg/kg (whole body) ⁴ 2,430 mg/kg (whole body) ⁴	Mortality, ED100 Mortality, ED100				[24] [24]	L; lifestage: 2-3 months L; lifestage: 2-3 months

Concentration units based on wet weight unless otherwise noted.
 BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

⁴ This entry was excerpted directly from the Environmental Residue-Effects Database (ERED, www.wes.army.mil/el/ered, U.S. Army Corps of Engineers and U.S. Environmental Protection Agency). The original publication was not reviewed, and the reader is strongly urged to consult the publication to confirm the information presented here.

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Chemical Category: PESTICIDE (ORGANOCHLORINE)

Chemical Name (Common Synonyms): DICOFOL CASRN: 115-32-2

Chemical Characteristics

Solubility in Water: 1.2 mg/L at 20° C **Half-Life:** No data [2]

(99% purity) [1]

 $Log K_{ow}$: No data [3] $Log K_{oc}$: —

Human Health

Oral RfD: 1 x 10⁻³ mg/kg/day [4] Confidence: —

Critical Effect: Increase in liver to body weight ratios in rats

Oral Slope Factor: No data [5] Carcinogenic Classification: C [6]

Wildlife

Partitioning Factors: Partitioning factors for dicofol in wildlife were not found in the literature.

Food Chain Multipliers: Food chain multipliers for dicofol in wildlife were not found in the literature.

Aquatic Organisms

Partitioning Factors: Log BCFs ranging from 4.02-4.16 were reported in a study exposing fathead minnows to dicofol [10].

Food Chain Multipliers: Food chain multipliers for dicofol in aquatic organisms were not found in the literature.

Toxicity/Bioaccumulation Assessment Profile

Dicofol is an organochlorine compound used as a miticide. The principal commercial dicofol product, Kelthane, is made from DDT [7]. Clark et al. [7] reported reduction in eggshell weight and thickness of American kestrels due to dicofol. They also observed that $10 \mu g/g$ of dicofol reduced hatchability of eggs. They suggested that dicofol concentrations above $3 \mu g/g$ in food may affect bird reproduction. The 48-h and 100-h LC50s for grass shrimp (*Crangon franciscorum*) exposed to dicofol (Kelthane) were 590 and $100 \mu g/L$, respectively [8].

The major metabolite of dicofol is 1,1-bis(4-chlorophenol)2,2dichloroethanol (pp-DCD) [9]. Because dicofol is more lipophilic than its metabolites, it was abundant in every tissue except for liver and brain. The dicofol metabolites are less toxic than dicofol and they have less impact on the formation of normal eggshells by doves [9]. The bioconcentration of dicofol in fathead minnows was reduced by 35 percent by clay particles (65 mg/L) indicating that more than 30 percent of the dicofol sorbed onto clay and was biologically unavailable to the fish [10]. Bioconcentration factors at the two dicofol concentrations were not significantly different and steady-state concentrations occured with 40 to 60 days of exposure at 10,500 to 13,900 times water levels.

Species:	Concentration, Units in ¹ :			Toxicity:	Ability to	Accumul	late ² :	Source:		
Taxa	Sediment	Water	Tissue (Sample Type)	Effects	Log BCF	Log BAF	BSAF	Reference	Comments ³	
Fishes										
Pimephales promelas, Fathead minnow		12.38 μg/L			4.02-4.16			[10]	L	
		1.15 µg/L			4.12-4.14			[10]	L	
Wildlife										
Streptopelia risoria, Ring neck dove	32 mg/kg (diet)		116.5μg/g in fat 1.07μg/g in liver 4.55μg/g in heart 0.37μg/g in brain					[9]	L	

¹ Concentration units based on wet weight unless otherwise noted.
² BCF = bioconcentration factor, BAF = bioaccumulation factor, BSAF = biota-sediment accumulation factor.

³ L = laboratory study, spiked sediment, single chemical; F = field study, multiple chemical exposure; other unusual study conditions or observations noted.

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